

The impact of repowering of wind farms on birds and bats



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1 Introduction

Wind energy generation has developed in Germany into by far the most important source of renewable energy production. Over 18,000 wind turbines with an installed capacity of over 19,000 MW are now in operation (status 30.6.2006, Bundesverband Windenergie, <http://windenergie.de/de/statistiken/>). The wind-rich offshore areas are seen as the best opportunity to increase the generation of wind energy. On the mainland, however, the expansion of wind energy is becoming increasingly more difficult, as most of the best sites are already taken and further extension of existing wind farms is restricted by the lack of wind in most of the inland regions, as well as planning restrictions for interests (protection of the environment, nature conservation and landscape). Therefore, „repowering“ provides a possibility to increase the production of electricity without simultaneously increasing the space required. „Repowering“ means that older, smaller and less powerful wind turbines are replaced by newer and more powerful ones. These new turbines are much higher than older ones - heights above 100m are now the norm.

The literature review of Hötker et al. (2005) (hereafter called the NABU-BfN report), which was funded by the German Agency for Nature Conservation (BfN), showed that wind turbines have only a relatively small disturbance effect on breeding birds, although many potentially sensitive species have not yet been analysed. Resting birds, in particular geese, ducks and waders responded sensitively to wind turbines and could be displaced from their resting areas. Wind turbines at certain sites, especially on bare mountain ridges and water bodies, are a collision risk for birds. In particular, birds of prey were affected, in Germany notably Red Kite and White-tailed Eagle. Bats were also killed by wind turbines, especially when these were located close to or within woodlands. The extent to which wind turbines have a harmful impact on the natural environment is mainly determined by their location. Practically the only effective means of avoiding damage was appropriate site selection.

The references used by Hötker et al. (2005) were generally studies which had been carried out on older, relatively small wind turbines. Although one could estimate a relationship between the size of wind turbines and the disturbance impact or collision risk for birds and bats, these relationships were not (with some exceptions) statistically significant. In order better to assess the impacts of the new generation of wind turbines on birds and bats, this report will analyse the more recent publications (up to summer 2006). The main emphasis was in particular the extent to which the danger for birds and bats caused by wind turbines is expected to change due to repowering. In this report, collisions of birds and bats and the displacement of birds by wind turbines have been considered as the potential impacts on these taxa. To our knowledge, no further essential studies about the displacement of bats and other mammals by wind turbines have been carried out recently (Bach & Rahmel, 2004) and therefore this aspect is not dealt with in this report.

2 Material and Methods

The methods used in this report are the same as the ones of the NABU-BfN report, so that the collected material of the NABU-BfN report could be included here. In addition, ca. 60 new publications were analysed, which had been derived from 45 individual studies. Data from each wind farm was treated as a single study, even if the data were gathered in different years and by different observers, to ensure the independence of the data and to avoid analysing the same study more than once.

The following publications were used in the NABU-BfN-report:

Ahlén, 2002; Albouy et al., 1997; Albouy et al., 2001; Anderson et al., 2000; Bach et al., 1999; Bach, 2001; Bach, 2002; Barrios & Rodriguez, 2004; Bergen, 2001a; Bergen, 2001b; Bergen, 2002a; Bergen, 2002b; Bergh et al., 2002; Boone, 2003; Böttger et al., 1990; Brauneis, 1999; Brauneis, 2000; Clemens & Lammen, 1995; De Lucas et al., 2004; Dulas Engineering Ltd, 1995; EAS, 1997; Erickson et al., 2003; Everaert, 2003; Everaert et al., 2002; Förster, 2003; Gerjets, 1999; Gharajedaghi & Ehrlinger, 2001; Guillemette & Larsen, 2002; Guillemette et al., 1999; Hall & Richards, 1962; Hormann, 2000; Hydro Tasmania; Isselbächer & Isselbächer, 2001; Janss, 2000; Johnson, 2002; Johnson et al., 2003; Johnson et al., 2000; Kaatz, 2000; Kaatz, 2002; Kerlinger, 2000; Ketzenberg et al., 2002; Koop, 1997; Koop, 1999; Korn & Scherner, 2000; Kowallik & Borbach-Jaene, 2001; Kruckenberg & Borbach-Jaene, 2001; Kruckenberg & Jaene, 1999; Leddy et al., 1999; Lekuona, 2001; Meek et al., 1993; Menzel, 2002; Menzel & Pohlmeier, 1999; Musters et al., 1996; Orloff & Flannery, 1996; Osborn et al., 1996; Pedersen & Poulsen, 1991; Percival, 2000; Phillips, 1994; Reichenbach, 2002; Reichenbach, 2003a; Reichenbach & Schadek, 2003; Reichenbach & Sinning, 2003; Sachslehner & Kollar, 1997; Scherner, 1999; Schmidt et al., 2003; Schreiber, 1992; Schreiber, 1993a; Schreiber, 1993b; Schreiber, 1999; Schreiber, 2002; SEO, 1995; SGS Environment, 1994; Sinning, 1999; Sinning & Gerjets, 1999; Smallwood & Thelander, 2004; Sommerhage, 1997; Steiof et al., 2002; Still et al., 1996; Strickland et al., 2001; Stübing & Bohle, 2001; Thelander & Rugge, 2000; Thelander et al., 2003; Trapp et al., 2002; van der Winden et al., 1999; Vierhaus, 2000; Walter & Brux, 1999; Winkelman, 1989; Winkelman, 1992a; Winkelman, 1992b; Young et al., 2003a; Young et al., 2003b.

And the following sources were included in this report:

Behr & Helversen, 2005; Brandt et al., 2005a; Brinkmann & Schauer-Weisshahn, 2006; Everaert & Stienen, 2006; Grünkorn et al., 2005; Handke et al., 2004a, b, c, d; Kerns et al., 2005; Koford et al., 2003; Lucas et al., 2005; Petersen et al., 2004; Reichenbach & Steinborn, 2006; Sinning, 2004a, b, c; Sinning & Bruyn, 2004; Sinning et al., 2004; Traxler et al., 2004.

As with the NABU-BfN report, the intention was to produce a report primarily relevant to Germany, and therefore the main emphasis was on research from Germany. The number of studies included from other countries also partly reflects here the scale of the research carried out in each country (Tab.1).

Country	Number of studies
Belgium	8
Germany	107
Denmark	3
France	2
Great Britain	6
Netherlands	5
Austria	5
Spain	11
USA	31
Australia	2

Table 1: Countries of the 180 studies evaluated in this report.

Most of the new studies cover several bird or bat species. For each species often several parameters were analysed (e.g. minimum distance to wind turbines and change in resting populations after installation of a wind farm; for further details see below). The division of species and parameters led to a data matrix with 207 data sets. Combined with the 1,789 data sets of Hötker et al. (2005), in total 1,996 data sets were available.

Most of the new data sets are quantitative analyses and only a few are „single observations“. Many of these „single observations“ have their source in systematic surveys, in the framework of which certain bird species were observed only rarely.

In spite of the addition of new references, the data material was still not suitable for a formal meta-analysis (Fernandez-Duque & Valeggia, 1994). Therefore, as with the NABU-BfN report, all the available results were included in the analysis. No distinction was made between results derived from systematic surveys and those based on only a few casual observations. The disadvantage of using all available studies is that, in statistical terms, casual observations are given equal weight to extensive research. It cannot be ruled out that „extreme“ observations have been recorded more frequently than less spectacular events. Also, additional factors, which could have been important in individual cases, may not have been fully considered. However, the advantage of this method is that the number of studies included in the analysis is large and therefore the results are less dependent on single, well-researched, but possibly atypical, studies. The independence of the data is also guaranteed. With a large number of studies, there is a greater chance that confounding factors may cancel each other out.

Unless stated otherwise, the statistical tests in this report use the „null-hypothesis“ that wind turbines have no influence on the parameter under consideration (for example population size before and after wind farm installation). The alternative hypothesis is that wind turbines do have an influence. In order to carry out the statistical tests, it was determined how many studies had a negative or positive effect (e.g. decrease or increase of population). As mentioned above, neither the strength of effect, nor statistical significance were considered. Neutral results (e.g. constant population) were classified as positive, in order to avoid any false association of wind energy with negative impacts. At the same time, statistically significant negative effects are made in this way more convincing and safer, and are not „diluted“ by the inclusion of neutral results. If wind power has no influence on bird populations, one would expect roughly equal proportions of positive and negative

effects. If the frequency of positive and negative effects differs strongly, some impact of wind energy can be assumed. In these cases, the statistical test used is the binomial test. Because not all of the available information is used in this procedure (for example the strength of the effects), it is very conservative, meaning that differences and trends are only classified as significant when they are very strong. The statistical tests were carried out using SPSS 7.5 statistical software.

Because the individual bird and bats species differ greatly in their biology and their use of habitat, whenever possible the evaluation was carried out for separate species. In cases when such differentiation was not possible, species were grouped.

It is assumed that animals which are relatively tied to their breeding areas react differently to wind turbines than those, which only pass through areas outside the breeding season, when they are less tied to a single area and lack local knowledge. Therefore a distinction was made between studies carried out during and outside the breeding season (definition varies, depending on the species concerned). Most studies did not indicate what activities the animals were carrying out at the time of observation (e.g. foraging, resting, roosting) and so this factor could not be considered in this report.

3 Analysed wind turbines

A substantial objective of this report was to estimate the impact that larger „new generation“ wind turbines have on birds and bats, as most of the data from the studies used by Hötker et al. (2005) was based on small wind turbines. For this reason, first of all the wind turbines used in this report shall be described and their different characteristics related to each other.

This report covers almost the whole spectrum of operational wind turbines used, from the beginning of wind energy production until the middle of 2006. The power capacity of the turbines ranges from less than 0.1 MW up to 2.0 MW. There was a similar spread for turbine hub height (22m to 114m), rotor diameter (14m to 80m) and accordingly total height (30m to 146m) to blade tip. As expected, the parameters capacity, hub height, rotor diameter and total height are closely correlated (Fig. 1-3). The relationship between the power capacity of the wind turbines and the remaining parameters can be described by the following equation, whereas the choice of the regression equation had been chosen in a way that the defining R² value was maximised (n=741):

$$\text{Hub height (m)} = 28.98 * \text{power capacity (MW)} + 30.29 \\ R^2 = 0.67 \text{ (p}<\text{0.001)}$$

$$\text{Total height (m)} = 87.01 * \text{power capacity (MW)}^{0.382} \\ R^2 = 0.73 \text{ (p}<\text{0.001)}$$

$$\text{Rotor diameter (m)} = 54.75 * \text{power capacity (MW)}^{0.382} \\ R^2 = 0.79 \text{ (p}<\text{0.001)}$$

In these equations and in Fig. 1-3, all available data points were used. Single wind turbines were thus counted repeatedly if more data analyses were available. This procedure was chosen because, first of all, it was necessary to clarify the relationships between the individual parameters within the data material, rather than to describe the technical development of the wind energy production.

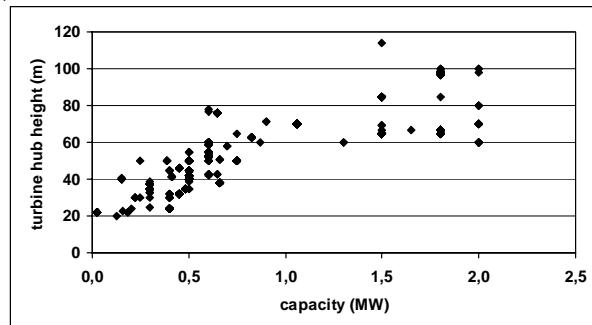


Figure 1: The relationship between power capacity of wind turbines and their hub height.

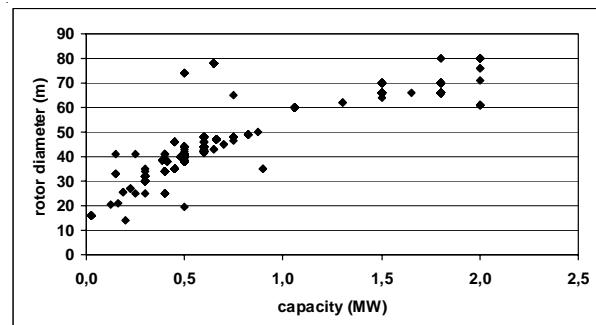


Figure 2: Relationship between the power capacity of wind turbines and their rotor diameter.

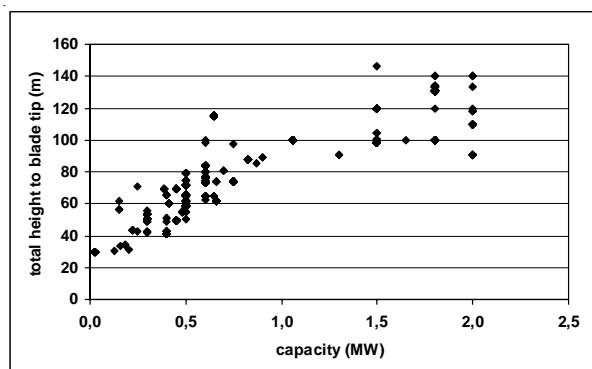


Figure 3: Relationship between the power capacity and the total height of wind turbines.

Power capacity and turbine size define strong linear trends in the early stages of technical development. However, from a capacity of around 1 MW, wind turbine size remains nearly constant as capacity increases.

Within the framework of this study, it was assumed that the overall height of a wind turbine is the parameter, most likely to determine if it has a disturbance impact and therefore in this report, the total height of turbines was the main parameter analysed.

4 Impacts of wind power on birds

4.1 Non-lethal impacts of wind turbines on birds (disturbance, displacement, habitat loss)

4.1.1 Change in distribution due to wind turbines

Because more material was available for this report than for the NABU-BfN report, the question of whether wind turbines have an impact on bird populations could be tested more thoroughly. Only activity taking place on the ground or in vegetation has been considered (breeding, resting, foraging). Despite the better availability of data, only a small number of studies permitted before-after-control impact comparison, and therefore studies that compare bird populations on the wind farm site with bird populations at similar sites in the surrounding area were also taken into account. As the analysed studies differ greatly from each other, it was only considered whether wind turbines had „positive“ or „negative“ effects. Negative effects are: (1) population decline after installation of the wind turbines; or (2) reduced numbers of birds within wind farms or the surrounding area, in comparison to control areas. Positive effects are accordingly population increase after installation of turbines, or increased bird numbers around the wind farm. The strength of the effects was not considered. If no population differences were detected, the effects were classified as positive in order to avoid falsely inflating the impact of negative effects (see above).

If wind energy has no impact, equal ratios of positive or negative effects are expected and statistical significance of these expectations was tested using a binomial test (for which the null hypothesis is that data are randomly distributed) (Tab. 2).

For 52 species or species groups enough studies (at least six) were of good enough quality to be included in statistical tests. Negative population impacts of wind farms during the breeding season could not be verified for any bird species. Only quail, redshank and lapwing displayed reduced numbers in connection with wind farms. The summarised material of all analysed studies on waders showed a statistically significant majority of negative reactions towards wind turbines. Positive or neutral effects predominate for the remaining species. Two species, which breed in reeds (marsh warbler and reed bunting) as well as stonechat even showed significantly more positive or neutral reactions towards wind turbines than negative ones.

One of the reasons for this phenomenon (i.e. „positive effects“) could be that due to construction of wind farms and their supply roads, trenches and high shrub fields had developed, which had not existed before in the uniform farmland and meadows. It is unlikely that these particular species are especially attracted to wind turbines themselves.

Studies carried out outside the breeding season show a very different picture and negative impacts of wind turbines predominate. Wigeon, Lapwing, Common Snipe and Golden Plover display significantly more negative than positive effects. The same applies to wildfowl which have been summarised according to their habits as followed: Geese (White-fronted Goose, Bean Goose, Greylag Goose, Brent Goose and Barnacle Goose), dabbling ducks (apart from Wigeon: Pintail, Shove-

Table 2: Impacts of wind turbines on bird populations showing the number of studies with positive or negative effects as revealed from literature (for details see text). The last column gives the result of sign tests (ns: not significant). Grey shading indicates predominantly negative effects.

Breeding season		Positive effect	Negative effect	Significance
Carrion/Hooded Crow	<i>Corvus corone</i>	6	2	ns
All Waders		30	53	0,016
Blackbird	<i>Turdus merula</i>	6	4	ns
Oystercatcher	<i>Haematopus ostralegus</i>	6	8	ns
White Wagtail	<i>Motacilla alba</i>	4	4	ns
Blue Tit	<i>Parus caeruleus</i>	4	3	ns
Whinchat	<i>Saxicola rubetra</i>	2	7	ns
Chaffinch	<i>Fringilla coelebs</i>	2	4	ns
Common Whitethroat	<i>Sylvia communis</i>	8	5	ns
Skylark	<i>Alauda arvensis</i>	18	16	ns
Willow Warbler	<i>Phylloscopus trochilus</i>	4	2	ns
Yellowhammer	<i>Emberiza citrinella</i>	4	6	ns
Linnet	<i>Carduelis cannabina</i>	3	6	ns
Northern Lapwing	<i>Vanellus vanellus</i>	12	23	ns
Common Buzzard	<i>Buteo buteo</i>	3	3	ns
Grey Partridge	<i>Perdix perdix</i>	5	5	ns
Reed Bunting	<i>Emberiza schoeniclus</i>	11	2	0,022
Redshank	<i>Tringa totanus</i>	2	9	ns
Yellow Wagtail	<i>Motacilla flava</i>	8	3	ns
Sedge Warbler	<i>Acrocephalus schoenobaenus</i>	10	0	0,002
Stonechat	<i>Saxicola torquata</i>	8	1	0,039
Mallard	<i>Anas platyrhynchos</i>	7	6	ns
Marsh Warbler	<i>Acrocephalus palustris</i>	7	4	ns
Reed Warbler	<i>Acrocephalus scirpaceus</i>	7	1	ns
Black-tailed Godwit	<i>Limosa limosa</i>	5	7	ns
Quail	<i>Coturnix coturnix</i>	1	6	ns
Meadow Pipit	<i>Anthus pratensis</i>	16	8	ns
Wren	<i>Troglodytes troglodytes</i>	6	1	ns
Chiffchaff	<i>Phylloscopus collybita</i>	4	2	ns
Outside breeding season				
Carrion/Hooded Crow	<i>Corvus corone</i>	13	8	ns
Oystercatcher	<i>Haematopus ostralegus</i>	4	3	ns
Common Snipe	<i>Gallinago gallinago</i>	0	6	0,05
Skylark	<i>Alauda arvensis</i>	5	2	ns
Geese		2	12	0,013
Golden Plover	<i>Pluvialis apricaria</i>	8	23	0,012
Grey Heron	<i>Ardea cinerea</i>	5	1	ns
Curlew	<i>Numenius arquata</i>	13	19	ns
Dabbling Ducks (except Wigeon)		3	15	0,008
Northern Lapwing	<i>Vanellus vanellus</i>	13	30	0,015
Black-headed Gull	<i>Larus ridibundus</i>	15	5	0,041
Common Buzzard	<i>Buteo buteo</i>	13	12	ns
Wigeon	<i>Anas penelope</i>	0	9	0,004
Tufted Duck	<i>Aythya fuligula</i>	2	6	ns
Wood Pigeon	<i>Columba palumbus</i>	2	7	ns
Red Kite	<i>Milvus milvus</i>	3	4	ns
Swans		2	6	ns
Herring Gull	<i>Larus argentatus</i>	2	5	ns
Starling	<i>Sturnus vulgaris</i>	17	6	0,035
Mallard	<i>Anas platyrhynchos</i>	3	8	ns
Common Gull	<i>Larus canus</i>	3	6	ns
Diving Ducks		2	12	0,013
Kestrel	<i>Falco tinnunculus</i>	15	7	ns
Fieldfare	<i>Turdus pilaris</i>	1	6	ns

ler, Mallard and Gadwall) and diving ducks (Common Pochard, Tufted Duck, Greater Scaup and Goldeneye). Exceptions were Starling and Black-headed Gull for which significantly more positive (or neutral) effects were recorded.

Overall, this study statistically confirms the results of others (Horch & Keller, 2005; Langston & Pullan, 2003; Reichenbach, 2003b), namely that wind farms have less impact on breeding birds, but greater impact on non-breeding birds. However, in contrast to the hypothesis so far, this analysis also showed that breeding populations may be locally threatened by wind turbines.

4.1.2 Minimum avoidance distance of birds to wind turbines

One of the essential questions of this report was to determine the impact particularly of newer, larger wind turbines on birds. It could be that disturbance factors of such turbines are much greater than those of smaller turbines previously analysed. Minimum avoidance distance of birds to wind turbines was included in 730 data sets. Results were available for 29 species or species groups or 26 species or species groups from the breeding season and the non-breeding season respectively (with at least four results for each species or species group to demonstrate a significant correlation of the chosen parameter in the analysis). The data are summarised in Table 3. With the help of regression calculations (the models used „power functions“), it was also possible to estimate to which extent the turbine influence displacement distances. Some of the studies are the same as those used in the previous chapter („impacts of wind turbines on population“).

The data show great variation, which was demonstrated by comparing results „between“ species and „within“ species. Therefore, in some cases standard deviations are very high (Tab.3). This may be explained either by the inclusion of casual observations that naturally show higher dispersion, or by larger differences between individual wind farms.

Despite the high degree of variation, some previously known trends could be clearly confirmed. Avoidance distances during the breeding season were smaller than outside the breeding season. Only a few wader species avoided close contact with wind farms at all times of the year.

Greater avoidance distances from wind turbines were generally observed outside the breeding season. As expected, birds of open habitats, e.g. geese, ducks and waders, generally avoided turbines by several hundred metres. However, under certain circumstances some individuals from nearly all of these species or species groups were observed very close to wind turbines. These observations were exceptional and therefore do not prove that the species concerned are generally insensitive to disturbance by wind turbines. In contrast, grey heron, birds of prey, oystercatcher, gulls, starling and crows were often observed close to or even within wind farms, which contributed partly to higher collision rates (see chapter 4.2.1 and Appendix). Sensitive species roosted at least 400-500m away from wind turbines (Tab.3). Greater avoidance distances are likely to cause negative effects only in exceptional circumstances. To a large extent, the results correspond with the conclusions of single studies on this topic (Kruckenberg & Jaene, 1999; Reichenbach, 2003b; Schreiber, 1993a; Schreiber, 1999).

Table 3: Minimum distances of different bird species from wind turbines according to different literature studies. The equations ($\text{minimum distance} = \text{coefficient} * \text{wind turbine height}^{\text{exponent}}$) demonstrate the relationship between the total height of wind turbines and the minimum distance of birds. Grey shading indicates cases where distance has increased with turbine height. SD: standard deviation. F: values of the variance-analyses in order to verify the regression coefficients.

Breeding season									
	Species	n	Median	Mean	SD	Coefficient	Exponent	p	F
Blackbird	Turdus merula	5	100	82	76	108812	-0,2290	ns	
Oystercatcher	Haematopus ostralegus	9	50	81	106	638,605	-0,7214	ns	
White Wagtail	Motacilla alba	5	50	72	51	5324,82	-1,0720	ns	
Bluethroat	Luscinia svecica	8	25	63	92	0,0004	2,5573	ns	
Curlew	Numenius arquata	4	125	163	144	83000000	-3,3837	ns	
Whinchat	Saxicola rubetra	5	125	155	60	125,073	0,0378	ns	
Common Whitethroat	Sylvia communis	12	70	75	57	852,679	-0,6752	ns	
Skylark	Alauda arvensis	26	105	120	116	3814,46	-0,9397	ns	
Finches		12	125	104	64	1659,37	-0,7498	ns	
Willow Warbler	Phylloscopus trochilus	5	50	42	40	59,5263	-0,0269	ns	
Garden Warbler	Sylvia borin	4	55	72	83	2179,7	-1,0076	ns	
Icterine Warbler	Hippolais icterina	4	30	40	45	372,29	-0,6644	ns	
Yellowhammer	Emberiza citrinella	6	85	89	58	149062	-1,6940	ns	
Corn Bunting	Miliaria calandra	4	88	94	88	1773392	-2,3029	ns	
Linnet	Carduelis cannabina	6	138	138	27	90,2427	0,0968	ns	
Northern Lapwing	Vanellus vanellus	21	125	134	119	0,3942	1,1575	ns	
Grey Partridge	Perdix perdix	4	100	125	96	9634801	-2,5701	ns	
Reed Bunting	Emberiza schoeniclus	16	50	86	139	2614,95	-1,0739	ns	
Redshank	Tringa totanus	6	188	183	111	1186,16	-0,4883	ns	
Yellow Wagtail	Motacilla flava	11	50	111	141	501,227	-0,5251	ns	
Sedge Warbler	Acrocephalus schoenobaenus	10	25	45	76	0,4507	0,9194	ns	
Stonechat	Saxicola torquata	5	50	104	150	50852,4	-1,5484	ns	
Starling	Sturnus vulgaris	4	75	71	62	4571,61	-1,0901	ns	
Mallard	Anas platyrhynchos	10	113	133	123	29,6433	0,2527	ns	
Marsh Warbler	Acrocephalus palustris	13	50	67	64	0,0032	2,1691	ns	
Reed Warbler	Acrocephalus scirpaceus	13	50	62	69	4288,03	-1,1862	ns	
Black-tailed Godwit	Limosa limosa	7	250	369	315	7826198	-2,3505	ns	
Meadow Pipit	Anthus pratensis	13	50	82	100	1111,88	-0,7998	ns	
Wren	Troglodytes troglodytes	5	50	90	96	0,0053	2,1329	ns	
Chiffchaff	Phylloscopus collybita	5	50	42	40	59,5263	-0,0269	ns	
Non-breeding season									
Carrion/Hooded Crow	Corvus corone	17	0	77	139	5E-09	5,0093	0,033	5,66
Oystercatcher	Haematopus ostralegus	6	15	55	81	3293811	-2,8716	ns	
Common Snipe	Gallinago gallinago	6	325	394	199	911,611	-0,2126	ns	
Coot	Fulica atra	4	138	136	99	1424,8	-0,6019	ns	
Curlew	Numenius arquata	25	200	222	178	236,007	-0,1474	ns	
Skylark	Alauda arvensis	6	0	38	59	0,0021	1,9466	ns	
Finches		14	45	58	59	1,6E-08	4,9391	<0,001	214,39
Geese		15	300	347	230	0,577	1,4018	ns	
Golden Plover	Pluvialis apricaria	24	150	202	190	0,004	3,0760	<0,001	21,14
Grey Heron	Ardea cinerea	7	60	120	170	3739,06	-1,0940	ns	
Northern Lapwing	Vanellus vanellus	36	175	273	390	0,000055	3,4002	<0,001	30,66
Black-headed Gull	Larus ridibundus	16	0	91	205	0,0114	1,7282	ns	
Common Buzzard	Buteo buteo	17	100	76	93	0,6489	0,9307	ns	
Gulls		32	25	120	208	0,3189	1,0722	ns	
Wigeon	Anas penelope	9	300	311	163	661,776	-0,2093	ns	
Wood Pigeon	Columba palumbus	6	100	175	178	4,9E-08	4,7582	ns	
Swans		8	125	150	139	5,4086	0,6210	ns	
Herring Gull	Larus argentatus	5	200	285	323	41,4305	0,2309	ns	
Starling	Sturnus vulgaris	18	0	38	58	0,000033	2,9925	0,036	5,4
Mallard	Anas platyrhynchos	9	200	161	139	1987,79	-0,8288	ns	
Common Gull	Larus canus	7	100	118	139	2,1054	0,7213	ns	
Diving Ducks		12	213	219	122	111,351	0,0673	ns	
Kestrel	Falco tinnunculus	16	0	36	53	2,2685	0,4728	ns	

When assessing the results, it should be remembered that some potentially sensitive species have still only been analysed rarely or not at all. This is particularly the case for the more controversial species (storks, birds of prey, cranes and corncrake) and therefore the list of species sensitive to disturbance is not complete.

As mentioned above, wind farms vary greatly with regard to their impacts on bird populations. It appears that the size of the turbine is at least partly responsible for these differences. The question of how turbine size affects minimum avoidance distances of birds to turbines is also relevant to repowering.

For bird species for which minimum avoidance distances were observed at least four different wind farms (the minimum number needed to obtain a statistically significant result), the relationship between turbine height and minimum avoidance distance was calculated and is presented in Table 3. This report could for the first time (compared with the NABU-BfN report) also include substantially more data relating to larger wind turbines.

Even though the remaining results presented in Tab. 3 are in most cases not statistically significant, overall this analysis still confirms other previous conclusions. Breeding birds were less disturbed by larger wind turbines than by smaller ones. 21 out of 29 species tended to use habitat closer to larger turbines than smaller ones. This was also the case for rather sensitive wader species, such as Black-tailed Godwit, Curlew and Redshank.

Outside the breeding season 16 out of 23 cases showed an increase in minimum avoidance distance with increased turbine size. For Lapwing, Golden Plover, Carrion Crow, Starling and Finches the results were statistically significant (Fig. 4-8).

In seven cases a negative correlation was noticeable, meaning that larger wind turbines had a smaller displacement effect. This also applies to species or species groups, which are more sensitive to disturbance, such as Wigeon, Common Snipe and Curlew.

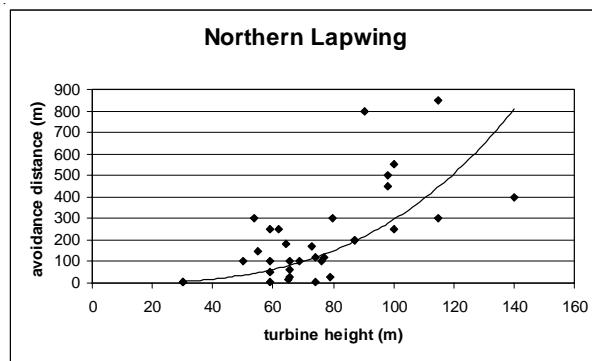


Figure 4: Relationship between minimum distance kept by Lapwings from wind turbines outside the breeding season in relation to total turbine height.

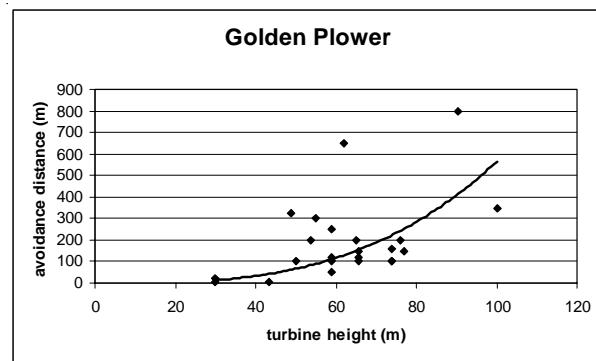


Figure 5: Relationship between minimum distance kept by Golden Plover from wind turbines outside the breeding season in relation to total turbine height.

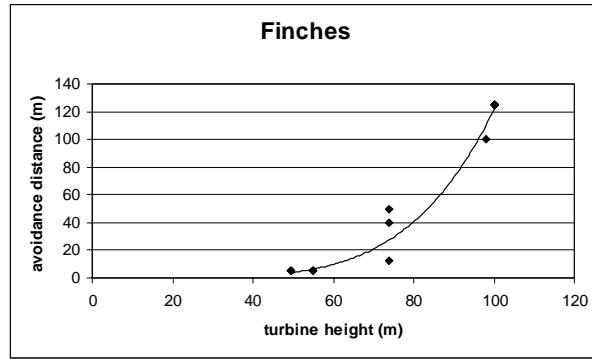


Figure 6: Relationship between minimum distance kept by Finches from wind turbines outside the breeding season in relation to total turbine height.

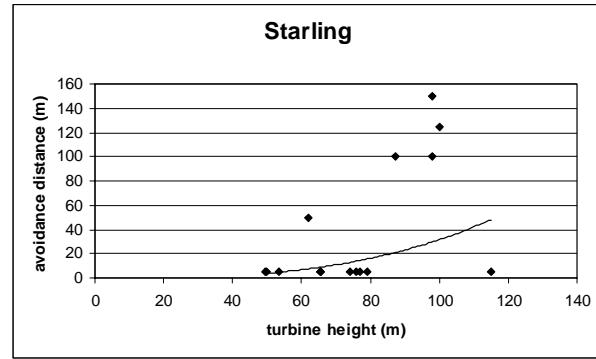


Figure 7: Relationship between minimum distance kept by Starling from wind turbines outside the breeding season in relation to total turbine height.

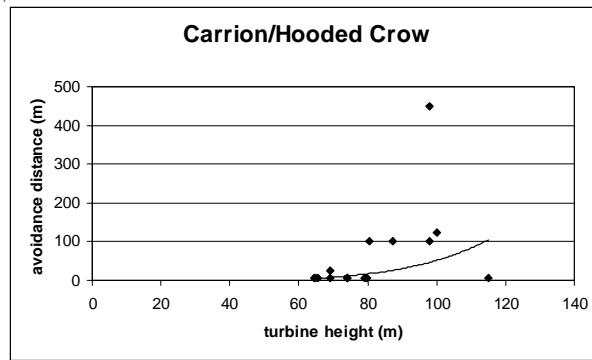


Figure 8: Relationship between minimum distance kept by Carrion/Hooded Crow from wind turbines outside the breeding season in relation to total turbine height.

4.2 Collision of birds and bats with wind turbines

4.2.1 Collision of birds with wind turbines

The NABU-BfN report included mainly observations available from the USA, so it was possible for this report to add some studies from Europe, which had been gathered particularly at relatively large and new wind turbines. Data were only included in the analysis if there were regular controls and if account was taken of the likelihood that carcasses disappeared from the study area (Anderson et al., 1999; Morrison, 2002). The mortality rates of Grünkorn et al. (2005) do not cover the whole year, but were still included in the analysis. Therefore, the data in Tab. 4 tend to underestimate rather than overestimate actual collision rates.

Some wind farms were represented in several studies, so that the present data sets partly overlap with each other. To guarantee the independence of the data sets, each wind farm was included only once in the statistical analysis, using either the most recent report, or that including the most extensive observations.

Collision rates varied greatly between different wind farms. For some wind farms no collisions or nearly none occurred. At other wind farms, collisions occurred at a frequency of more than 60 per turbine per year. Mass collisions, similar to those known from lighthouses or other buildings (Crawford & Engstrom, 2001; Erickson et al., 2002; Manville, 2001; Ugoretz, 2001) could not be identified for individual turbines within wind farms. In most of the studies the casualty rate is one bird per turbine per year; the median was 1.8 and the mean 6.9 victims per turbine per year.

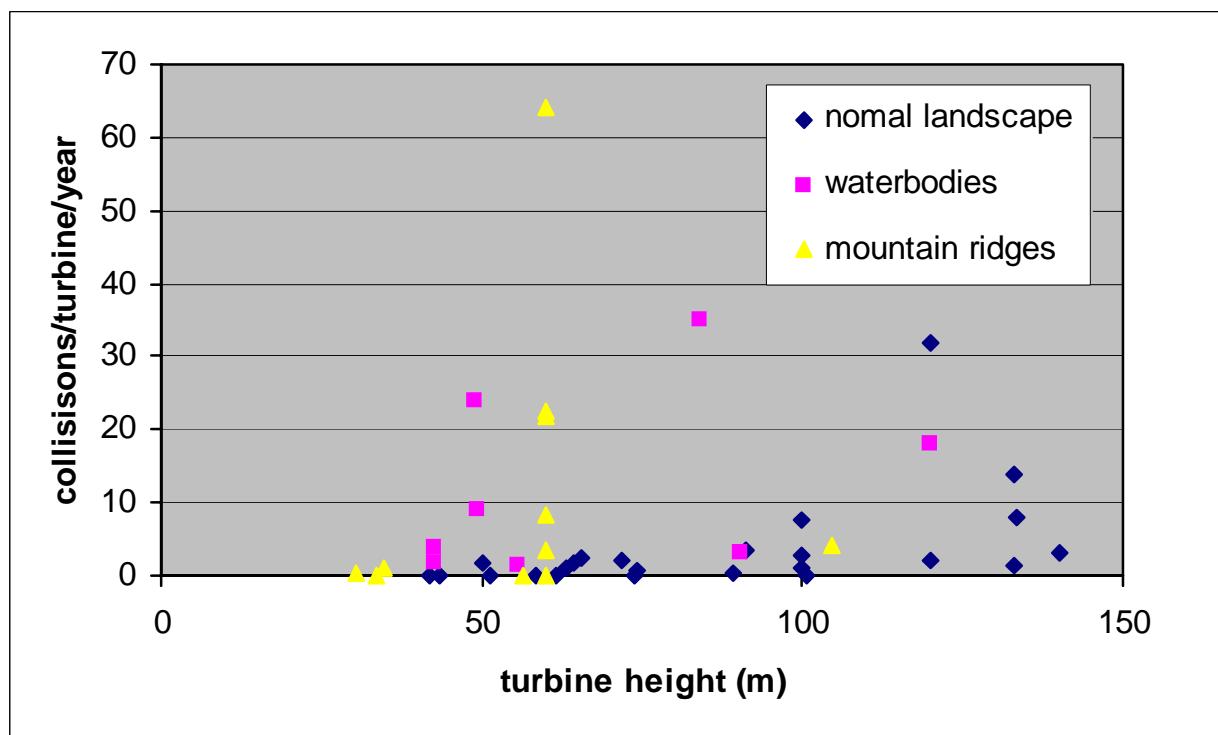


Figure 9: Collision rates of birds with wind turbines at different wind farm sites in relation to total turbine height.

Table 4: Collision rates birds (mean number of victims per turbine per year) at different wind farms.

Country	Wind farm	Habitat	Collisions / Turbine / Year	Comments	Source
Australia	Tasmania	Coast	1,86		Hydro Tasmania
Belgium	Boudewijnkanaal te Brugge	Wetland	35	Other studies in different years	Everaert et al., 2003
Belgium	Elektriciteitscentrale te Schelle	Wetland	18	Other studies in different years	Everaert et al., 2003
Belgium	Gent	Industrial	2		Everaert in litt
Belgium	Kleine Pathoekeweg, Brugge	Industrial	32		Everaert in litt
Belgium	Kluizendok, Gent	Industrial	8		Everaert in litt
Belgium	Nieuwkapelle, Diksmuide	Wet grassland	1		Everaert in litt
Belgium	Oostdam te Zeebrugge	Wetland	24	Other studies in different years	Everaert, Devos & Kuijken, 2003
Denmark	Tjaereborg	Wetland	3		Pedersen & Poulsen, 1991b
Germany	Breklumer Koog	Fields	>7,5	Study period less than 1 year	Grünkorn et al, 2005
Germany	Bremerhaven-Fischereihafen	Wetland	9		Scherner, 1999b
Germany	Friedrich-W ilhelm-Lübke-Koog	Fields	>2,6	Study period less than 1 year	Grünkorn et al, 2005
Germany	Simonsberger Koog	Fields	>2,2	Study period less than 1 year	Grünkorn et al, 2005
Netherlands	Kreekrak sluice	Wetland	3,7		Musters et al., 1996
Netherlands	Oosterbierum	Grassland	1,8		Winkelman, 1992a
Netherlands	Urk	Coast	1,7		W inkelman, 1989
Austria	Obersdorf	Woodland edge, fields	1,49		Traxler et al., 2005
Austria	Prellenkirchen	Fields	13,93		Traxler et al., 2006
Austria	Steinberg-Prinzendorf	Woodland edge, fields	2,99		Traxler et al., 2004
Sweden	Näsudden	Grassland	0,7		Percival, 2000
Spain	Alaiz-Echague	Mountain ridges	3,56		Lekuona, 2001
Spain	E3, Energia Eólica del Estrecho	Mountain ridges	0,03		Barrios & Rodriguez, 2004; SEO, 1995
Spain	El Perdón	Mountain ridges	64,26		Lekuona, 2001
Spain	Guennda	Mountain ridges	8,47		Lekuona, 2001
Spain	Izco-Albar	Mountain ridges	22,63		Lekuona, 2001
Spain	PESUR, Parque Eólico del Sur and Parque und Parque Eólico de	Mountain ridges	0,36		Barrios & Rodriguez, 2004; SEO, 1995
Spain	Levantera				
Spain	Salajones	Mountain ridges	21,69		Lekuona, 2001
Spain	Tarifa		0,03		Janss, 2000
UK	Blyth	Wetland	1,34		Still et al., 1996
UK	Bryn Tytli	Moor, grassland	0		Phillips, 1994
UK	Burgar Hill, Orkney	Moor, grassland	0,15		Percival, 2000
UK	Cemmaes	Moor, grassland	0,04		Percival, 2000
UK	Haverigg, Cumbria	Moor, grassland	0		Percival, 2000
UK	Ovenden Moor	Moor, grassland	0,04		Percival, 2000
USA	Altamont	Mountain ridges	0,87	Other studies in different years	Smallwood & Thelander, 2004
USA	Buffalo Ridge	Grassland	0,98	Other studies in different years	Erickson et al., 2001
USA	Foote Creek Rim	Prairie	1,75	Other studies in different years	Erickson et al., 2001
USA	Green Mt, Searsburg	Mountain ridges	0		Erickson et al., 2001
USA	IDWGP, Algona	Mountain ridges	0		Erickson et al., 2001
USA	Mountaineer	Mountain ridges	4,04		Kerns & Kerlinger, 2004
USA	Nine Canyon W ind Project	Prairie	3,59	Other studies in different years	Erickson et al., 2003
USA	San Gorgino	Mountain ridges	2,31		Erickson et al., 2001
USA	Solano County	Mountain ridges	54		Erickson et al., 2001
USA	Somerset County	Mountain ridges	0		Erickson et al., 2001
USA	Top of Iowa	Prairie	0,415		Koford et al., 2003
USA	Vansycle	Fields, grassland	0,63	Other studies in different years	Erickson et al., 2001

One of the main emphases of this report was the issue of the extent to which collision rates were dependent on turbine size. The worry was that relatively more birds are killed by particularly large wind turbines than at smaller ones. Collision rates correlate significantly with hub and total height (Fig. 9, Tab.5).

It is already known from previous analyses that collision rates were particularly high at certain locations, such as bare mountain ridges and water bodies. It was also possible to confirm these results in this report. Giving equal weight to „habitat“ (category high risk: mountain ridges and water bodies; category low risk: other habitats) and „turbine height“, a GLM-analysis identified both habitat and turbine height as significant ($F_1 = 7.96$; $p = 0.007$) or nearly significant impact parameters ($F_1 = 3.37$; $p = 0.074$), respectively. Wind farms located in low-risk „normal“ landscapes show a relationship between collision rate and the size of wind turbines (Fig. 9) that is statistically significant (Tab.5). If considering the turbine size dependency of bird casualties within the two habitat-types wetlands and normal landscape, there was no significant result in either case.

In order to address the issue of what species are particularly affected by wind turbines, the comprehensive statistical analyses of T. Dürr (*Staatliche Vogelschutzwarte Brandenburg*) should be pointed out, and the latest version was included in this report. The species composition of collision victims has not fundamentally changed. In Germany, birds of prey dominate with very high numbers of Red Kites and White-tailed Eagles.

4.2.2 Collision of bats with wind farms

Table 5: The relationship between collision rates of birds and hub height, rotor diameter and total height of wind turbines. In addition, the relationship between collision rates of birds and total turbine height in different habitat types is included.

Type	Parameter	n	R ²	Regression equation	F	p
All data	Hub height	43	0,110	$Y = 0,0006 x^{1,948}$	0,431	0,030
All data	Rotor diameter	43	0,084	$Y = 0,0007 x^{1,966}$	3,75	0,06
All data	Total height	43	0,105	$Y = 0,0002 x^{2,022}$	4,78	0,034
“Normal” landscape	Total height	24	0,360	$Y = 0,00000017 x^{3,978}$	12,39	0,002
Wetlands	Total height	8	0,167	$Y = 0,0303 x^{1,302}$	1,20	0,32
Mountain ridges	Total height	11	0,146	$Y = 0,0000032 x^{3,240}$	1,53	0,25

It has been known for some time that bats can also be killed by wind turbines. More recent studies (Brinkmann & Schauer-Weisshahn, 2006) have improved the quality of data. As for the NABU-BfN report, data were only analysed if the number of victims per turbine per year could be calculated. As with the bird analyses, account was also taken of the possibility that bat carcasses could have been removed by scavengers. Because bats were mainly killed in summer and early autumn, only observations from this time period were included. In order to ensure the independence of the data, only one value per wind farm was used, as for the bird analyses.

Collision rates for bats (Tab.6) show even more variation than collision rates for birds (Tab.4). Even though at many wind farms no bats or only a few were killed,

other wind farms showed high mortality rates. The results for 34 wind farms ranged between 0 and 134 victims (median: 6.4; mean: 13.3; standard deviation: 13.3 for killed bats per turbine per year), and the highest rates exceed the highest rates for birds.

Table 6: Collision rates of bats (average number of victims per turbine per year) at different wind farms.

Country	Windfarm	Habitat	Collisions / Turbine / Year	Source
Australia	Tasmania	Coast	1,86	Hydro Tasmania
Germany	Ettenheim Brudergarten 1	Woodland	35,18	Brinkmann & Schauer Weisshahn, 2005
Germany	Ettenheim Brudergarten 2	Woodland	24,12	Brinkmann & Schauer Weisshahn, 2005
Germany	Ettenheim Brudergarten 3	Woodland/ windthrow	22,04	Brinkmann & Schauer Weisshahn, 2005
Germany	Ettenheim Mahlberg 1	Windthrow	13,02	Brinkmann & Schauer Weisshahn, 2005
Germany	Ettenheim Mahlberg 2	Woodland	9,62	Brinkmann & Schauer Weisshahn, 2005
Germany	Ettenheim Mahlberg 3	Woodland	14,64	Brinkmann & Schauer Weisshahn, 2005
Germany	Freiamt Hohe Eck	Woodland	52,34	Brinkmann & Schauer Weisshahn, 2005
Germany	Freiamt Schillinger Berg 1	Woodland	103,16	Brinkmann & Schauer Weisshahn, 2005
Germany	Freiamt Schillinger Berg 2	Meadow	0	Brinkmann & Schauer Weisshahn, 2005
Germany	Fürstenberg	Meadow	0	Brinkmann & Schauer Weisshahn, 2005
Germany	Horben Holzschlägermatte 1	Woodland	37,56	Brinkmann & Schauer Weisshahn, 2005
Germany	Horben Holzschlägermatte 2	Woodland	8,02	Brinkmann & Schauer Weisshahn, 2005
Germany	Rosskopf	Woodland, mountain ridges	21,5	Behr & Helversen, 2005
Germany	Simonswald Plattenhöfe 2	Meadow/ woodland	7,59	Brinkmann & Schauer Weisshahn, 2005
Germany	Simonswald Plattenhöfe 3	Meadow/ woodland	7,94	Brinkmann & Schauer Weisshahn, 2005
Germany	Simonswald Plattenhöfe 4	Meadow	0	Brinkmann & Schauer Weisshahn, 2005
Germany	St. Peter Plattenhöfe 1	Meadow/ woodland	0	Brinkmann & Schauer Weisshahn, 2005
Austria	Obersdorf	Woodland edge, fields	0	Traxler et al., 2005
Austria	Prellenkirchen	Fields	8	Traxler et al., 2005
Austria	Steinberg-Prinzendorf	Woodland edge, fields	5,33	Traxler et al., 2005
Spain	Alaiz-Echague	Mountain ridges	0	Lekuona, 2001
Spain	El Perdón	Mountain ridges	0	Lekuona, 2001
Spain	Guennda	Mountain ridges	0	Lekuona, 2001
Spain	Izco-Albar	Mountain ridges	3,09	Lekuona, 2001
Spain	Salajones	Mountain ridges	13,36	Lekuona, 2001
USA	Altamont	Mountain ridges	0,0035	Smallwood & Thelander, 2004
USA	Buffalo Ridge	Grassland	2,3	Osborn et al., 1996
USA	Foote Creek Rim	Prairie	1,34	Young et al., 2003a
USA	Mautaineer Wind Energy Facility Blackwater Falls	Woodland	50	Boone, 2003
USA	Meyersdale	Woodland, mountain ridges	25	Kerns et al., 2005
USA	Mountaineer	Woodland, mountain ridges	38	Kerns & Kerlinger, 2004
USA	Nine Canyon Wind Project	Prairie	3,21	Erickson et al., 2003
USA	Top of Iowa	Woodland, mountain ridges	6,432	Koford et al., 2003
USA	Vansycle	Fields, grassland	0,4	Strickland et al., 2001b

Furthermore, there was a statistically significant relationship between the number of bats killed and hub height, rotor diameter and wind turbine height (Fig. 10, Tab. 7). However, if one takes into account the fact that bats are killed more frequently at sites close to woodlands, then the influence of turbine size disappears. A GLM-analysis shows that the factor „habitat“ (woods and other habitats) had an almost significant influence on the collision rates ($F_1 = 3.801$; $p = 0.06$), while turbine height showed no influence ($F_1 = 0.17$; $p = 0.69$). Considering turbine height and collision rates separately for sites close to woodlands and in other locations, then no significant relationship is noticeable (Tab.7).

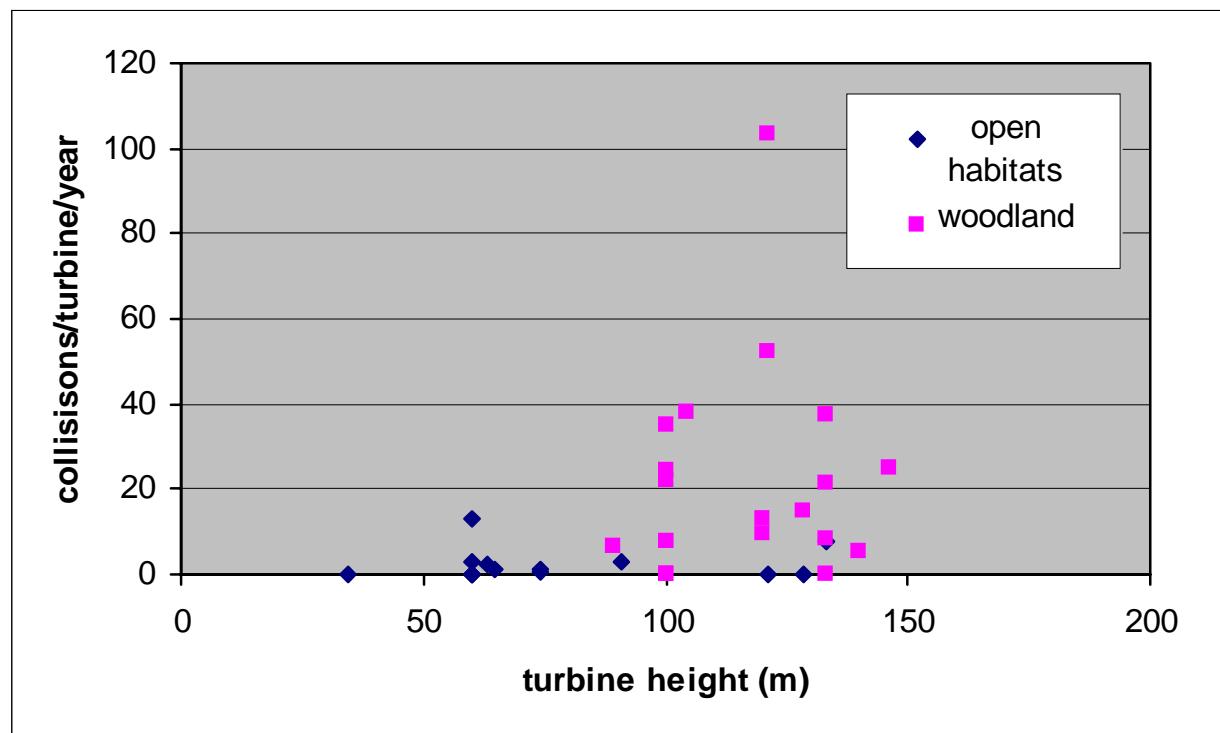


Figure 10: Collision rates of bats with wind turbines of different total heights.

Table 7: Relationship between collision rates of bats and the hub height, rotor diameter and the total height of wind turbines at different wind farms. In addition, the relationship between the collision rate of birds and the total height of wind turbines in different habitat types is included.

Type	Parameter	n	R ²	Regression equation	F	p
All data	Hub height	34	0,139	$Y = 0,0000025 \times^{3,180}$	5,17	0,030
All data	Rotor diameter	34	0,165	$Y = 0,000000019 \times^{4,517}$	6,33	0,017
All data	Total height	34	0,152	$Y = 0,000000086 \times^{3,674}$	5,75	0,022
Open habitat	Total height	14	0,007	$Y = 0,0134 \times^{0,679}$	0,09	0,77
Woodland locations	Total height	20	0,022	$Y = 0,0000069 \times^{2,877}$	0,41	0,53

For further information on the species composition of collision victims, see T. Dürr/*Staatliche Vogelschutzwarte Brandenburg* (see Appendix). At 14 wind farms, collisions of both bats and birds were analysed and there was no relationship between collision rates ($R^2 = 0,009$, $p = 0,75$).

5 Estimated impacts of repowering

Increasing the amount of electricity produced by wind energy on land will be achieved through repowering, meaning that many small wind turbines will be replaced by larger, more modern and thus more powerful wind turbines. With the help of the results collected so far, this chapter will try to assess the impacts of repowering expected on birds and bats. Displacement impacts of repowering shall be considered, as well as the risk of collisions. Four scenarios were modelled as a first

step towards understanding the impacts of repowering. Scenarios 1 and 2 assume that 0.3 MW wind turbines are replaced by 1.5 MW and 2.0 MW turbines respectively, while in scenarios 3 and 4, 0.5 MW turbines are replaced by 1.5 MW and 2.0 MW wind turbines respectively. For each scenario, the model simulates various increases in capacity (ranging between no increase (factor 1) up to a five-times increase).

In order to assess impacts on birds and bats, first of all the expected turbine height was calculated using the relationship between turbine capacity and height (see chapter 3).

5.1 Repowering and disturbance of birds

The relationship between total turbine height and minimum distances birds are found from turbines, combined with the relationship between turbine height and capacity, allows the effects of repowering on the spatial distribution of birds to be estimated. Hence, the following simple hypotheses are assumed:

- 1 No birds use the area within the radius of the minimum distance (disturbance area); outside this radius turbines have no effect.
- 2 Scenarios consider single, standalone wind turbines. Because wind farms vary greatly in layout, effects cannot be generalised. The effect of layout must be considered separately for each wind farm.

Impacts can only be estimated by comparing the size of the disturbed areas. For example, at one wind farm 20 x 0.3MW wind turbines shall be replaced by 8 x 1.5MW turbines. The capacity of the wind farm increases by a factor of 2. The repowering has a negative effect for a bird species, if the sum of the disturbance factors from 8 x 1.5MW wind turbines is bigger than the sum of disturbance factors from 20 x 0.3MW turbines.

The disturbance influence radii of wind turbines of different heights can be calculated using the data in Table 3. Examples of model calculations are summarised in Table 8. The results of the different scenarios are not substantially different from each other. In most cases breeding birds reacted positively to repowering, while visiting birds showed a variable picture in which repowering had a negative effect on sensitive species, such as geese, Golden Plover and Lapwing.

Table 8: Estimates of disturbance factors for birds after repowering, using model calculations under different scenarios. Positive (+) means a smaller area of disturbance, while negative (-) means a larger area of disturbance after repowering. See text for more details.

Breeding season; Scenario: 0,3 MW to 1,5 MW

Species	Increase in output (factors):	1	1,5	2	2,5	3	3,5	4	4,5	5
Oystercatcher	<i>Haematopus ostralegus</i>	+	+	+	+	+	+	+	+	+
White Wagtail	<i>Motacilla alba</i>	+	+	+	+	+	+	+	+	+
Bluethroat	<i>Luscinia svecica</i>	-	-	-	-	-	-	-	-	-
Curlew	<i>Numenius arquata</i>	+	+	+	+	+	+	+	+	+
Whinchat	<i>Saxicola rubetra</i>	+	+	+	+	+	+	+	+	-
Common Whitethroat	<i>Sylvia communis</i>	+	+	+	+	+	+	+	+	+
Skylark	<i>Alauda arvensis</i>	+	+	+	+	+	+	+	+	+
Finches		+	+	+	+	+	+	+	+	+
Willow Warbler	<i>Phylloscopus trochilus</i>	+	+	+	+	+	+	+	+	+
Garden Warbler	<i>Sylvia borin</i>	+	+	+	+	+	+	+	+	+
Icterine Warbler	<i>Hippolais icterina</i>	+	+	+	+	+	+	+	+	+
Yellowhammer	<i>Emberiza citrinella</i>	+	+	+	+	+	+	+	+	+
Corn Bunting	<i>Miliaria calandra</i>	+	+	+	+	+	+	+	+	+
Linnet	<i>Carduelis cannabina</i>	+	+	+	+	+	+	+	-	-
Northern Lapwing	<i>Vanellus vanellus</i>	+	-	-	-	-	-	-	-	-
Grey Partridge	<i>Perdix perdix</i>	+	+	+	+	+	+	+	+	+
Reed Bunting	<i>Emberiza schoeniclus</i>	+	+	+	+	+	+	+	+	+
Redshank	<i>Tringa totanus</i>	+	+	+	+	+	+	+	+	+
Yellow Wagtail	<i>Motacilla flava</i>	+	+	+	+	+	+	+	+	+
Sedge Warbler	<i>Acrocephalus schoenobaenus</i>	+	+	-	-	-	-	-	-	-
Stonechat	<i>Saxicola torquata</i>	+	+	+	+	+	+	+	+	+
Starling	<i>Sturnus vulgaris</i>	+	+	+	+	+	+	+	+	+
Mallard	<i>Anas platyrhynchos</i>	+	+	+	+	+	+	-	-	-
Marsh warbler	<i>Acrocephalus palustris</i>	-	-	-	-	-	-	-	-	-
Reed warbler	<i>Acrocephalus scirpaceus</i>	+	+	+	+	+	+	+	+	+
Black-tailed godwit	<i>Limosa limosa</i>	+	+	+	+	+	+	+	+	+
Meadow pipit	<i>Anthus pratensis</i>	+	+	+	+	+	+	+	+	+
Wren	<i>Troglodytes troglodytes</i>	-	-	-	-	-	-	-	-	-
Chiffchaff	<i>Phylloscopus collybita</i>	+	+	+	+	+	+	+	+	+

Outside the breeding season; Scenario: 0,3 MW to 1,5 MW

Species	Increase in output (factors):	1	1,5	2	2,5	3	3,5	4	4,5	5
Carrión/Hooded Crow	<i>Corvus corone</i>	-	-	-	-	-	-	-	-	-
Blackbird	<i>Turdus merula</i>	+	+	+	+	+	+	+	+	+
Oystercatcher	<i>Haematopus ostralegus</i>	+	+	+	+	+	+	+	+	+
Common Snipe	<i>Gallinago gallinago</i>	+	+	+	+	+	+	+	+	+
Coot	<i>Fulica atra</i>	+	+	+	+	+	+	+	+	+
Curlew	<i>Numenius arquata</i>	+	+	+	+	+	+	+	+	+
Chaffinch	<i>Fringilla coelebs</i>	-	-	-	-	-	-	-	-	-
Skylark	<i>Alauda arvensis</i>	-	-	-	-	-	-	-	-	-
Finches		-	-	-	-	-	-	-	-	-
Geese		-	-	-	-	-	-	-	-	-
Golden Plover	<i>Pluvialis apricaria</i>	-	-	-	-	-	-	-	-	-
Grey Heron	<i>Ardea cinerea</i>	+	+	+	+	+	+	+	+	+
Northern Lapwing	<i>Vanellus vanellus</i>	-	-	-	-	-	-	-	-	-
Black-headed Gull	<i>Larus ridibundus</i>	-	-	-	-	-	-	-	-	-
Common Buzzard	<i>Buteo buteo</i>	+	+	-	-	-	-	-	-	-
Gulls		+	-	-	-	-	-	-	-	-
Wigeon	<i>Anas penelope</i>	+	+	+	+	+	+	+	+	+
Wood Pigeon	<i>Columba palumbus</i>	-	-	-	-	-	-	-	-	-
Swans		+	+	+	-	-	-	-	-	-
Herring Gull	<i>Larus argentatus</i>	+	+	+	+	+	+	-	-	-
Starling	<i>Sturnus vulgaris</i>	-	-	-	-	-	-	-	-	-
Mallard	<i>Anas platyrhynchos</i>	+	+	+	+	+	+	+	+	+
Common Gull	<i>Larus canus</i>	+	+	+	-	-	-	-	-	-
Diving Ducks		+	+	+	+	+	+	+	+	-
Kestrel	<i>Falco tinnunculus</i>	+	+	+	+	-	-	-	-	-

Breeding season; Scenario: 0,3 MW to 2,0 MW

Species	Increase in output (factors):	1	1,5	2	2,5	3	3,5	4	4,5	5
Oystercatcher	<i>Haematopus ostralegus</i>	+	+	+	+	+	+	+	+	+
White Wagtail	<i>Motacilla alba</i>	+	+	+	+	+	+	+	+	+
Bluethroat	<i>Luscinia svecica</i>	-	-	-	-	-	-	-	-	-
Curlew	<i>Numenius arquata</i>	+	+	+	+	+	+	+	+	+
Whinchat	<i>Saxicola rubetra</i>	+	+	+	+	+	+	+	+	+
Common Whitethroat	<i>Sylvia communis</i>	+	+	+	+	+	+	+	+	+
Skylark	<i>Alauda arvensis</i>	+	+	+	+	+	+	+	+	+
Finches		+	+	+	+	+	+	+	+	+
Willow Warbler	<i>Phylloscopus trochilus</i>	+	+	+	+	+	+	+	+	+
Garden Warbler	<i>Sylvia borin</i>	+	+	+	+	+	+	+	+	+
Icterine Warbler	<i>Hippolais icterina</i>	+	+	+	+	+	+	+	+	+
Yellowhammer	<i>Emberiza citrinella</i>	+	+	+	+	+	+	+	+	+
Corn Bunting	<i>Miliaria calandra</i>	+	+	+	+	+	+	+	+	+
Linnet	<i>Carduelis cannabina</i>	+	+	+	+	+	+	+	+	+
Northern Lapwing	<i>Vanellus vanellus</i>	+	-	-	-	-	-	-	-	-
Grey Partridge	<i>Perdix perdix</i>	+	+	+	+	+	+	+	+	+
Reed Bunting	<i>Emberiza schoeniclus</i>	+	+	+	+	+	+	+	+	+
Redshank	<i>Tringa totanus</i>	+	+	+	+	+	+	+	+	+
Yellow Wagtail	<i>Motacilla flava</i>	+	+	+	+	+	+	+	+	+
Sedge Warbler	<i>Acrocephalus schoenobaenus</i>	+	+	-	-	-	-	-	-	-
Stonechat	<i>Saxicola torquata</i>	+	+	+	+	+	+	+	+	+
Starling	<i>Sturnus vulgaris</i>	+	+	+	+	+	+	+	+	+
Mallard	<i>Anas platyrhynchos</i>	+	+	+	+	+	+	+	+	-
Marsh Warbler	<i>Acrocephalus palustris</i>	-	-	-	-	-	-	-	-	-
Reed Warbler	<i>Acrocephalus scirpaceus</i>	+	+	+	+	+	+	+	+	+
Black-tailed Godwit	<i>Limosa limosa</i>	+	+	+	+	+	+	+	+	+
Meadow Pipit	<i>Anthus pratensis</i>	+	+	+	+	+	+	+	+	+
Wren	<i>Troglodytes troglodytes</i>	-	-	-	-	-	-	-	-	-
Chiffchaff	<i>Phylloscopus collybita</i>	+	+	+	+	+	+	+	+	+

Outside the breeding season; scenario: 0,3 MW to 2,0 MW

Species	Increase in output (factors):	1	1,5	2	2,5	3	3,5	4	4,5	5
Carrion/Hooded Crow	<i>Corvus corone</i>	-	-	-	-	-	-	-	-	-
Blackbird	<i>Turdus merula</i>	+	+	+	+	+	+	+	+	+
Oystercatcher	<i>Haematopus ostralegus</i>	+	+	+	+	+	+	+	+	+
Common Snipe	<i>Gallinago gallinago</i>	+	+	+	+	+	+	+	+	+
Coot	<i>Fulica atra</i>	+	+	+	+	+	+	+	+	+
Curlew	<i>Numenius arquata</i>	+	+	+	+	+	+	+	+	+
Chaffinch	<i>Fringilla coelebs</i>	-	-	-	-	-	-	-	-	-
Skylark	<i>Alauda arvensis</i>	-	-	-	-	-	-	-	-	-
Finches		-	-	-	-	-	-	-	-	-
Geese		-	-	-	-	-	-	-	-	-
Golden Plover	<i>Pluvialis apricaria</i>	-	-	-	-	-	-	-	-	-
Grey Heron	<i>Ardea cinerea</i>	+	+	+	+	+	+	+	+	+
Northern Lapwing	<i>Vanellus vanellus</i>	-	-	-	-	-	-	-	-	-
Black-headed Gull	<i>Larus ridibundus</i>	-	-	-	-	-	-	-	-	-
Common Buzzard	<i>Buteo buteo</i>	+	+	-	-	-	-	-	-	-
Gulls		+	-	-	-	-	-	-	-	-
Wigeon	<i>Anas penelope</i>	+	+	+	+	+	+	+	+	+
Wood Pigeon	<i>Columba palumbus</i>	-	-	-	-	-	-	-	-	-
Swans		+	+	+	+	-	-	-	-	-
Herring Gull	<i>Larus argentatus</i>	+	+	+	+	+	+	+	+	-
Starling	<i>Sturnus vulgaris</i>	-	-	-	-	-	-	-	-	-
Mallard	<i>Anas platyrhynchos</i>	+	+	+	+	+	+	+	+	+
Common Gull	<i>Larus canus</i>	+	+	+	-	-	-	-	-	-
Diving Ducks		+	+	+	+	+	+	+	+	+
Kestrel	<i>Falco tinnunculus</i>	+	+	+	+	+	-	-	-	-

Breeding season; scenario: 0,5 MW to 1,5 MW

Species	Increase in output (factors):	1	1,5	2	2,5	3	3,5	4	4,5	5
Oystercatcher	<i>Haematopus ostralegus</i>	+	+	+	+	+	+	+	+	+
White Wagtail	<i>Motacilla alba</i>	+	+	+	+	+	+	+	+	+
Bluethroat	<i>Luscinia svecica</i>	-	-	-	-	-	-	-	-	-
Curlew	<i>Numenius arquata</i>	+	+	+	+	+	+	+	+	+
Whinchat	<i>Saxicola rubetra</i>	+	+	+	+	-	-	-	-	-
Common Whitethroat	<i>Sylvia communis</i>	+	+	+	+	+	+	+	+	+
Skylark	<i>Alauda arvensis</i>	+	+	+	+	+	+	+	+	+
Finches		+	+	+	+	+	+	+	+	+
Willow Warbler	<i>Phylloscopus trochilus</i>	+	+	+	+	+	-	-	-	-
Garden Warbler	<i>Sylvia borin</i>	+	+	+	+	+	+	+	+	+
Icterine Warbler	<i>Hippolais icterina</i>	+	+	+	+	+	+	+	+	+
Yellowhammer	<i>Emberiza citrinella</i>	+	+	+	+	+	+	+	+	+
Corn Bunting	<i>Miliaria calandra</i>	+	+	+	+	+	+	+	+	+
Linnet	<i>Carduelis cannabina</i>	+	+	+	+	-	-	-	-	-
Northern Lapwing	<i>Vanellus vanellus</i>	+	-	-	-	-	-	-	-	-
Grey Partridge	<i>Perdix perdix</i>	+	+	+	+	+	+	+	+	+
Reed Bunting	<i>Emberiza schoeniclus</i>	+	+	+	+	+	+	+	+	+
Redshank	<i>Tringa totanus</i>	+	+	+	+	+	+	+	-	-
Yellow Wagtail	<i>Motacilla flava</i>	+	+	+	+	+	+	+	+	-
Sedge Warbler	<i>Acrocephalus schoenobaenus</i>	+	-	-	-	-	-	-	-	-
Stonechat	<i>Saxicola torquata</i>	+	+	+	+	+	+	+	+	+
Starling	<i>Sturnus vulgaris</i>	+	+	+	+	+	+	+	+	+
Mallard	<i>Anas platyrhynchos</i>	+	+	+	-	-	-	-	-	-
Marsh Warbler	<i>Acrocephalus palustris</i>	-	-	-	-	-	-	-	-	-
Reed Warbler	<i>Acrocephalus scirpaceus</i>	+	+	+	+	+	+	+	+	+
Black-tailed Godwit	<i>Limosa limosa</i>	+	+	+	+	+	+	+	+	+
Meadow Pipit	<i>Anthus pratensis</i>	+	+	+	+	+	+	+	+	+
Wren	<i>Troglodytes troglodytes</i>	-	-	-	-	-	-	-	-	-
Chiffchaff	<i>Phylloscopus collybita</i>	+	+	+	+	+	-	-	-	-

Outside the breeding season; scenario: 0,5 MW to 1,5 MW

Species	Increase in output (factors):	1	1,5	2	2,5	3	3,5	4	4,5	5
Carrion/Hooded Crow	<i>Corvus corone</i>	-	-	-	-	-	-	-	-	-
Blackbird	<i>Turdus merula</i>	+	+	+	+	+	+	-	-	-
Oystercatcher	<i>Haematopus ostralegus</i>	+	+	+	+	+	+	+	+	+
Common Snipe	<i>Gallinago gallinago</i>	+	+	+	+	+	+	-	-	-
Coot	<i>Fulica atra</i>	+	+	+	+	+	+	+	+	-
Curlew	<i>Numenius arquata</i>	+	+	+	+	+	-	-	-	-
Chaffinch	<i>Fringilla coelebs</i>	-	-	-	-	-	-	-	-	-
Skylark	<i>Alauda arvensis</i>	-	-	-	-	-	-	-	-	-
Finches		-	-	-	-	-	-	-	-	-
Geese		+	-	-	-	-	-	-	-	-
Golden Plover	<i>Pluvialis apricaria</i>	-	-	-	-	-	-	-	-	-
Grey Heron	<i>Ardea cinerea</i>	+	+	+	+	+	+	+	+	+
Northern Lapwing	<i>Vanellus vanellus</i>	-	-	-	-	-	-	-	-	-
Black-headed Gull	<i>Larus ridibundus</i>	+	+	-	-	-	-	-	-	-
Common Buzzard	<i>Buteo buteo</i>	+	-	-	-	-	-	-	-	-
Gulls		+	-	-	-	-	-	-	-	-
Wigeon	<i>Anas penelope</i>	+	+	+	+	+	+	-	-	-
Wood Pigeon	<i>Columba palumbus</i>	-	-	-	-	-	-	-	-	-
Swans		+	+	-	-	-	-	-	-	-
Herring Gull	<i>Larus argentatus</i>	+	+	+	-	-	-	-	-	-
Starling	<i>Sturnus vulgaris</i>	-	-	-	-	-	-	-	-	-
Mallard	<i>Anas platyrhynchos</i>	+	+	+	+	+	+	+	+	+
Common Gull	<i>Larus canus</i>	+	-	-	-	-	-	-	-	-
Diving Ducks		+	+	+	+	-	-	-	-	-
Kestrel	<i>Falco tinnunculus</i>	+	+	+	-	-	-	-	-	-

Breeding season; scenario: 0,5 MW to 2,0 MW

Species	Increase in output (factors):	1	1,5	2	2,5	3	3,5	4	4,5	5
Oystercatcher	<i>Haematopus ostralegus</i>	+	+	+	+	+	+	+	+	+
White Wagtail	<i>Motacilla alba</i>	+	+	+	+	+	+	+	+	+
Bluethroat	<i>Luscinia svecica</i>	-	-	-	-	-	-	-	-	-
Curlew	<i>Numenius arquata</i>	+	+	+	+	+	+	+	+	+
Whinchat	<i>Saxicola rubetra</i>	+	+	+	+	+	+	-	-	-
Common Whitethroat	<i>Sylvia communis</i>	+	+	+	+	+	+	+	+	+
Skylark	<i>Alauda arvensis</i>	+	+	+	+	+	+	+	+	+
Finches		+	+	+	+	+	+	+	+	+
Willow Warbler	<i>Phylloscopus trochilus</i>	+	+	+	+	+	+	+	+	+
Garden Warbler	<i>Sylvia borin</i>	+	+	+	+	+	+	+	+	+
Icterine Warbler	<i>Hippolais icterina</i>	+	+	+	+	+	+	+	+	+
Yellowhammer	<i>Emberiza citrinella</i>	+	+	+	+	+	+	+	+	+
Corn Bunting	<i>Miliaria calandra</i>	+	+	+	+	+	+	+	+	+
Linnet	<i>Carduelis cannabina</i>	+	+	+	+	+	+	-	-	-
Northern Lapwing	<i>Vanellus vanellus</i>	+	-	-	-	-	-	-	-	-
Grey Partridge	<i>Perdix perdix</i>	+	+	+	+	+	+	+	+	+
Reed Bunting	<i>Emberiza schoeniclus</i>	+	+	+	+	+	+	+	+	+
Redshank	<i>Tringa totanus</i>	+	+	+	+	+	+	+	+	+
Yellow Wagtail	<i>Motacilla flava</i>	+	+	+	+	+	+	+	+	+
Sedge Warbler	<i>Acrocephalus schoenobaenus</i>	+	+	-	-	-	-	-	-	-
Stonechat	<i>Saxicola torquata</i>	+	+	+	+	+	+	+	+	+
Starling	<i>Sturnus vulgaris</i>	+	+	+	+	+	+	+	+	+
Mallard	<i>Anas platyrhynchos</i>	+	+	+	+	+	-	-	-	-
Marsh Warbler	<i>Acrocephalus palustris</i>	-	-	-	-	-	-	-	-	-
Reed Warbler	<i>Acrocephalus scirpaceus</i>	+	+	+	+	+	+	+	+	+
Black-tailed Godwit	<i>Limosa limosa</i>	+	+	+	+	+	+	+	+	+
Meadow Pipit	<i>Anthus pratensis</i>	+	+	+	+	+	+	+	+	+
Wren	<i>Troglodytes troglodytes</i>	-	-	-	-	-	-	-	-	-
Chiffchaff	<i>Phylloscopus collybita</i>	+	+	+	+	+	+	+	+	+

Outside the breeding season; scenario: 0,5 MW to 2,0 MW

Species	Increase in output (factors):	1	1,5	2	2,5	3	3,5	4	4,5	5
Carriion/Hooded Crow	<i>Corvus corone</i>	-	-	-	-	-	-	-	-	-
Blackbird	<i>Turdus merula</i>	+	+	+	+	+	+	+	+	+
Oystercatcher	<i>Haematopus ostralegus</i>	+	+	+	+	+	+	+	+	+
Common Snipe	<i>Gallinago gallinago</i>	+	+	+	+	+	+	+	+	+
Coot	<i>Fulica atra</i>	+	+	+	+	+	+	+	+	+
Curlew	<i>Numenius arquata</i>	+	+	+	+	+	+	+	+	-
Chaffinch	<i>Fringilla coelebs</i>	-	-	-	-	-	-	-	-	-
Skylark	<i>Alauda arvensis</i>	-	-	-	-	-	-	-	-	-
Finches		-	-	-	-	-	-	-	-	-
Geese		-	-	-	-	-	-	-	-	-
Golden Plover	<i>Pluvialis apricaria</i>	-	-	-	-	-	-	-	-	-
Grey Heron	<i>Ardea cinerea</i>	+	+	+	+	+	+	+	+	+
Northern Lapwing	<i>Vanellus vanellus</i>	-	-	-	-	-	-	-	-	-
Black-headed Gull	<i>Larus ridibundus</i>	-	-	-	-	-	-	-	-	-
Common Buzzard	<i>Buteo buteo</i>	+	+	-	-	-	-	-	-	-
Gulls		+	-	-	-	-	-	-	-	-
Wigeon	<i>Anas penelope</i>	+	+	+	+	+	+	+	+	+
Wood Pigeon	<i>Columba palumbus</i>	-	-	-	-	-	-	-	-	-
Swans		+	+	-	-	-	-	-	-	-
Herring Gull	<i>Larus argentatus</i>	+	+	+	+	+	-	-	-	-
Starling	<i>Sturnus vulgaris</i>	-	-	-	-	-	-	-	-	-
Mallard	<i>Anas platyrhynchos</i>	+	+	+	+	+	+	+	+	+
Common Gull	<i>Larus canus</i>	+	+	-	-	-	-	-	-	-
Diving Ducks		+	+	+	+	+	+	-	-	-
Kestrel	<i>Falco tinnunculus</i>	+	+	+	-	-	-	-	-	-

5.2 Repowering and collisions of birds and bats

Following a similar procedure to that in the previous section it is possible to calculate collision rates for the different scenarios, using the relationship between wind turbine capacity and turbine height, and turbine height and collision rates. The estimates presented in Tab. 9 are for wind turbines in „normal“ landscape. As shown in chapter 4.2, considerably higher collision rates of birds and bats would occur at wind farms located close to water bodies or woodlands, respectively. In both habitats, no wind farms should be installed or replaced by repowering and therefore these habitats were not considered further.

The results of modelling show that in all cases repowering has a negative impact on birds – larger wind turbines have higher collision rates than smaller ones (see also chapter 4.2).

Bats show a different picture. The relationship between wind turbine height and the number of casualties is very weak. The extent to which repowering has an impact depends on the factor by which the power capacity is increased. Up to a doubling in capacity, casualties should be less, while larger increases in capacity would result in increased mortality rates.

Table 9: Assessment of collision rates of birds and bats following repowering using model calculations. Positive impacts (+) mean a reduced collision risk while negative impacts (-) indicate increased risk. For more details see text.

Collision rates of birds

Changes in collision rates of birds with increased wind farm output (factors)

Scenario	1	1,5	2	2,5	3	3,5	4	4,5	5
0,3 MW to 1,5 MW	-	-	-	-	-	-	-	-	-
0,3 MW to 2,0 MW	-	-	-	-	-	-	-	-	-
0,5 MW to 1,5 MW	-	-	-	-	-	-	-	-	-
0,5 MW to 2,0 MW	-	-	-	-	-	-	-	-	-

Collision rates of bats

Changes in collision rates of bats with increased wind farm output (factors)

Scenario	1	1,5	2	2,5	3	3,5	4	4,5	5
0,3 MW to 1,5 MW	+	+	+	-	-	-	-	-	-
0,3 MW to 2,0 MW	+	+	+	-	-	-	-	-	-
0,5 MW to 1,5 MW	+	+	-	-	-	-	-	-	-
0,5 MW to 2,0 MW	+	+	-	-	-	-	-	-	-

6 Discussion and research requirements

Newly published results regarding the issue of birds, bats and wind farms allow a better understanding of the impacts of larger, more modern wind turbines, which could not be considered previously (e.g. NABU-BfN report (Hötker et al. 2005)). Many previous results were confirmed by this study. It seems that wind turbines have a small impact on breeding birds, apart from particularly sensitive species (waders). However, there is still not much data available to show how other potentially sensitive species, or species of particular nature conservation interest (e.g. large birds and most bird of prey) react to wind turbines.

Wind turbines displace birds from their resting and feeding areas outside the breeding season. This has been confirmed for ducks, geese and some wader species. No new findings have been published regarding minimum displacement distances of birds from wind turbines. Starling and finches, but not Lapwings, showed a significant relationship between turbine height and avoidance distance. The new generation of wind turbines does not have a more disturbing effect on breeding birds than smaller less powerful turbines. This is due to the fact that increases in energy efficiency are proportionally greater than increases in turbine size alone. Among other things, birds, in particular breeding birds, seem to be displaced less by larger turbines than by smaller or middle-sized ones, in contrast to visiting birds, which were more sensitive to larger wind turbines. There was no change to the list of species previously classified as sensitive; the comprehensive case studies were confirmed. However, the particular noticeable results of Brandt et al. (2005b) at the Wybelsumer Polder are in strong contrast to other publications. Unfortunately, the presentation of data in that report did not allow a further analysis and could not be included in the database.

With respect to collision rates, the assumption that the choice of the wind farm site has a significant influence on the impacts of wind turbines was confirmed. High casualty rates for birds occur at wind farms on bare mountain ridges and water bodies, and close to woodlands for bats. Mass mortalities were still not reported for migrating birds. Taking into account the influence of habitat, it can be shown that a clear relationship exists for birds between turbine height and collision rate. A wind turbine 100m high is responsible on average for five bird casualties per year. Despite the better data, no statistically significant relationship between turbine height and mortality rate for bats could be found. Habitat choice is the dominant influence for bats.

Most of the results used in this report are based on only a small number of studies: this applies to avoidance distances as well as to collision rates, and therefore further research is necessary to confirm the results to date. In particular, no data has been published so far which studies the collision rates at large wind turbines at well-known migration hotspots. The numerous existing wind farms in Schleswig-Holstein on migration routes could help to determine whether wind farms have an impact on migrating birds.

Studies analysing the behaviour of large birds particularly birds of prey at wind farms are urgently needed. Birds of prey are strongly affected (see Appendix) and in Germany Red Kites and White-tailed Eagles have especially high number of

casualties. Using either automatic recording equipment or targeted observations, it should be possible to determine the situations in which these species are particularly threatened, and thus the collision rates of red kites and sea eagles could hopefully be reduced in the long term.

As there are many wind farms in Schleswig-Holstein, particularly on the grazing marshes at the west coast, it is necessary to study the extent to which the presence of wind farms may already have influenced the habitats of species occupying these areas.

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8 Summary

The impact of repowering of wind farms on birds and bats

The references used by Hötker et al. (2005) in the NABU-BfN report and 45 new studies were analysed in order better to assess the impacts of large, more modern wind turbines on both birds and bats.

This analysis basically confirmed the results of the NABU-BfN report. Wind turbines have small impacts on breeding birds, apart from waders, which were significantly displaced by wind turbines during the breeding season. However, wind turbines have very large impacts on migrant birds outside the breeding season and for ducks, geese and some other waders species a displacement effect could statistically be proven.

Estimates of avoidance distances of larger wind turbines have not changed as a result of this analysis. Birds of open habitats (especially ducks, geese and some waders species) in most cases kept an avoidance distance of several hundred metres from wind turbines. Lapwing, Golden Plover, Carrion Crow, Starling and finches showed significant relationships between turbine height and avoidance distance outside the breeding season. During the breeding season, however, it was not possible to determine whether new generation wind turbines have greater disturbance effects than smaller, less powerful turbines.

Collision rates for birds and bats depend on the choice of wind farm site. Wind farms located close to water bodies or bare mountain ridges were responsible for significantly more bird casualties than others, while wind farms located close to woodlands showed high collision rates for bats. Collision rates at these sensitive sites were many times higher than at other less sensitive sites, so that, for examp-

Ie, at some wind turbines more than 100 bats were killed per year. Taking habitat choice into account, there was a significant relationship between collision risk and wind turbine height for birds, but not for bats.

With regard to repowering, the results imply that possible disturbance effects need to be assessed differently, depending on the range of bird species occurring on a particular wind farm site. For most breeding birds repowering has a positive effect, while outside the breeding season further species-specific evaluations are necessary, as repowering might increase the collision risk for some birds. If the capacity of a wind turbine is more than doubled, then collision rates for bats should also increase. The differences are, however, minor.

This report also identifies gaps in knowledge, as well as future research requirements.

9 References

- Ahlén, I. (2002). Fladdermöss och fåglar dödade av vindkraftverk. *Fauna och Flora*, 97, 14-21.
- Albouy, S., Clément, D., Jonard, A., Massé, P., Pagès, J.-M., & Neau, P. (1997). Suivi ornithologique du parc éolien de Porte-la-Nouvelle (Aude) - Rapport final. ABIES, LPO, Gardouch.
- Albouy, S., Dubois, Y., & Picq, H. (2001). Suivi ornithologique des parcs éoliens du plateau de Garrigue Haute (Aude) - Rapport final. ABIES, LPO, Gardouch.
- Anderson, R.L., Strickland, M.D., Tom, J., Neumann, N., Erickson, W.P., Cleckler, J., Mayorga, G., Nuhn, G., Leuders, A., Schneider, J., Backus, L., Becker, P., & Flagg, N. (2000). Avian Monitoring and Risk Assessment at Tehachapi Pass and San Gorgonio Pass Wind Resource Areas, California. In: Proceedings of National Avian - Wind Power Planning Meeting III (ed PNAWPPM-III), pp. 31-46. Prepared for the Avian Subcommittee of the National Wind Coordinating Committee by LGL Ltd., King City, Ont., San Diego, California.
- Bach, L. (2001). Fledermäuse und Windenergienutzung - reale Probleme oder Einbildung? *Vogelkundliche Berichte aus Niedersachsen*, 33, 119-124.
- Bach, L. (2002). Auswirkungen von Windenergieanlagen auf das Verhalten von Fledermäusen am Beispiel des Windparks „Hohe Geest“, Midlum. Bericht der Arbeitsgemeinschaft zur Förderung angewandter biologischer Forschung im Auftrag der KW Midlum GmbH & Co. KG, Freiburg, Niederelbe.
- Bach, L., Handke, K., & Sinning, F. (1999). Einfluss von Windenergieanlagen auf die Verteilung von Brut- und Rastvögeln in Nordwest-Deutschland. *Bremer Beiträge für Naturkunde und Naturschutz*, 4, 107-122.
- Bach, L. & Rahmel, U. (2004). Überblick zu Auswirkungen von Windkraftanlagen auf Fledermäuse - eine Konfliktabschätzung. *Bremer Beiträge für Naturkunde und Naturschutz*, 7, 245-252.
- Barrios, L. & Rodriguez, A. (2004). Behavioural and environmental correlates of soaring bird mortality at on-shore wind turbines. *Journal of Applied Ecology*, 41, 72-81.
- Behr, O. & Helversen, O.V. (2005). Gutachten zur Beeinträchtigung im freien Luftraum jagender und ziehender Fledermäuse durch bestehende Windkraftanlagen - Wirkungskontrolle zum Windpark „Rosskopf“ (Freiburg i. Br.). Univ. Erlangen-Nürnberg, Inst. für Zoologie.

- Bergen, F. (2001a). Untersuchungen zum Einfluss der Errichtung und des Betriebs von Windenergieanlagen auf Vögel im Binnenland. Ph D thesis, Ruhr Universität, Bochum.
- Bergen, F. (2001b). Windkraftanlagen und Frühjahrsdurchzug des Kiebitz (*Vanellus vanellus*): eine Vorher/Nacher-Studie an einem traditionellen Rastplatz in Nordrhein-Westfalen. Vogelkundliche Berichte aus Niedersachsen, 33, 89-96.
- Bergen, F., ed. (2002a). Einfluss von Windenergieanlagen auf die Raum-Zeitnutzung von Greifvögeln, pp 86-96. Technische Universität Berlin, Berlin.
- Bergen, F., ed. (2002b). Windkraftanlagen und Frühjahrsdurchzug des Kiebitz (*Vanellus vanellus*): eine Vorher-Nachher-Studie an einem traditionellen Rastplatz in Nordrhein-Westfalen, pp 77-85. Technische Universität Berlin, Berlin.
- Bergh, L.M.J.v.d., Spaans, A.L., & Swelm, N.D.v. (2002). Lijnopstellingen van windturbines geen barrière voor voedselvluchten van meeuvens en sterns in de broedtijd. Limosa, 75, 25-32.
- Boone, D. (2003). Bat kill at West Virginia windplant, Maryland.
- Böttger, M., Clemens, T., Grote, G., Hartmann, G., Hartwig, E., Lammen, C., Vauk-Hentzelt, E., & Vauk, G., eds. (1990). Biologisch-Ökologische Begleituntersuchungen zum Bau und Betrieb von Windkraftanlagen. NNA.
- Brandt, U., Butenschön, S., Denker, E., & Ratzbor, G. (2005a). Brütend unterm Windrad - Entwicklung eines national bedeutsamen Brutvogelbiotops am Dollart.
- Brandt, U., Butenschön, S., Denker, E. & Ratzbor, G. (2005b). Rast am Rotor - Gastvogel-Monitoring im und am Windpark Wybelsumer Polder. UVP-Report 19: 170 - 174.
- Brauneis, W. (1999). Der Einfluss von Windkraftanlagen auf die Avifauna am Beispiel der „Solzer Höhe“ bei Bebra-Solz im Landkreis Hersfeld-Rotenburg. Untersuchung im Auftrag des Bundes für Umwelt und Naturschutz (BUND) Landesverband Hessen e.V. - Ortsverband Alheim-Rotenburg-Bebra, Bebra.
- Brauneis, W. (2000). Der Einfluss von Windkraftanlagen (WKA) auf die Avifauna, dargestellt insb. am Beispiel des Kranichs *Grus grus*. Ornithologische Mitteilungen, 52, 410-415.
- Brinkmann, R. & Schauer-Weisshahn, H. (2006). Untersuchungen zu möglichen betriebsbedingten Auswirkungen von Windkraftanlagen auf Fledermäuse im Regierungsbezirk Freiburg. Regierungspräsidium Freiburg, Stiftung Naturschutzfonds Baden-Württemberg, Grundelfingen.
- Clemens, T. & Lammen, C. (1995). Windkraftanlagen und Rastplätze von Küstenvögeln - ein Nutzungskonflikt. Seevögel, 16, 34-38.
- Crawford, R.L. & Engstrom, R.T. (2001). Characteristics of avian mortality at a north Florida television tower: a 29-year study. Journal of Field Ornithology, 72, 380-388.
- De Lucas, M., Janss, G.F.E., & Ferrer, M. (2004). The effects of a wind farm on birds in a migration point: the Strait of Gibraltar. Biodiversity and Conservation, 13, 395-407.
- Dulas Engineering Ltd (1995). The Mynydd y Cemmaes windfarm impact study. Vol. IID - Ecological impact - final report. ETSU report: W/13/00300/REP2D.
- EAS (1997). Ovenden Moor Ornithological Monitoring. Report to Yorkshire Windpower. Keighly: Ecological Advisory Service.
- Erickson, W., Johnson, G., Young, D., Strickland, D., Good, R., Bourassa, M., Bay, K., & Sernka, K.J. (2002). Synthesis and comparison of baseline avian and bat use, raptor nesting and mortality information from proposed and existing wind developments. Report for Bonneville Power Administration, Portland, Oregon.
- Erickson, W., Kronner, K., & Gritski, B. (2003). Nine Canyon Wind Power Project. Avian and Bat Monitoring Report. September 2002 - August 2003. Prepared for Nine Canyon Technical Advisory Committee by West, Inc., Cheyenne.

- Everaert, J. (2003). Collision victims on 3 wind farms in Flanders (Belgium) in 2002. Instituut voor Naturbeheer, Brussel.
- Everaert, J., Devos, K., & Kuijken, E. (2002). Windturbines en vogels in Vlaanderen, Rep. No. 2002.3. Instituut voor Natuurbehoud, Brussels.
- Everaert, J. & Stienen, E.W.M. (2006). Impact of wind turbines on birds in Zeebrugge (Belgium) - Significant effects on breeding tern colony due to collisions. *Biodiversity and Conservation*, in press.
- Fernandez-Duque, E. & Valeggia, C. (1994). Meta-analysis: a valuable tool in conservation research. *Conservation Biology*, 8, 555-561.
- Förster, F. (2003). Windkraft und Fledermausschutz in der Oberlausitz. In: Kommen die Vögel und Fledermäuse unter die (Wind)räder?, Dresden, 17.-18.11.2003.
- Gerjets, D. (1999). Annäherung wiesenbrütender Vögel an Windkraftanlagen - Ergebnisse einer Brutvogeluntersuchung im Nahbereich des Windparks Drochtersen. Bremer Beiträge für Naturkunde und Naturschutz, 4, 49-52.
- Gharadjedaghi, B. & Ehrlinger, M. (2001). Auswirkungen des Windparks bei Nitzschka (Lkr. Altenburger Land) auf die Vogelfauna. *Landschaftspflege und Naturschutz in Thüringen*, 38, 73-83.
- Grünkorn, T., Diederichs, A., Stahl, B., Poszig, D., & Nehls, G. (2005). Entwicklung einer Methode zur Abschätzung des Kollisionsrisikos von Vögeln an Windenergieanlagen. Bio Consult SH im Auftrag des Landesamts für Natur und Umwelt Schleswig Holstein, Hockensbüll.
- Guillemette, M. & Larsen, J.K. (2002). Postdevelopment experiments to detect anthropogenic disturbances: the case of sea ducks and wind parks. *Ecological Applications*, 12, 868-877.
- Guillemette, M., Larsen, J.K., & Clausanger, I. (1999). Assessing the impact of the Tunø Knob wind park on sea ducks: the influence of the food resources, Rep. No. Technical Report No. 263. National Environmental Research Institute, Denmark.
- Hall, L.S. & Richards, G.C. (1962). Notes on *Tadarida australis* (Chiroptera: molossidae). *Australian Mammalogy*, 1, 46.
- Handke, K., Adena, J., Handke, P., & Sprötge, M. (2004a). Einfluss von Windenergieanlagen auf die Verteilung ausgewählter Brut- und Rastvogelarten in einem Bereich der Krummhörn (Jennelt/Ostfriesland). Bremer Beiträge für Naturkunde und Naturschutz, 7, 47-58.
- Handke, K., Adena, J., Handke, P., & Sprötge, M. (2004b). Räumliche Verteilung ausgewählter Brut- und Rastvogelarten in Bezug auf vorhandene Windenergieanlagen in einem Bereich der küstennahen Krummhörn. Bremer Beiträge für Naturkunde und Naturschutz, 7, 11-44.
- Handke, K., Adena, J., Handke, P., & Sprötge, M. (2004c). Untersuchungen an ausgewählten Brutvogelarten nach Errichtung eines Windparks im Bereich der Stader Geest (Landkreis Rotenburg/Wümme und Stade). Bremer Beiträge für Naturkunde und Naturschutz, 7, 69-75.
- Handke, K., Adena, J., Handke, P., & Sprötge, M. (2004d). Untersuchungen zum Vorkommen von Kiebitz (*Vanellus vanellus*) und Großem Brachvogel (*Numenius arquata*) vor und nach Errichtung. Bremer Beiträge für Naturkunde und Naturschutz, 7, 61-66.
- Horch, P. & Keller, V. (2005). Windkraftanlagen und Vögel - ein Konflikt? Eine Literaturrecherche. Schweizerische Vogelwarte Sempach, Sempach.
- Hormann, M. (2000). Schwarzstorch - *Ciconia nigra*. In Avifauna von Hessen, 4. Lieferung. HGON.
- Hötker, H., Thomsen, K.-M., & Köster, H. (2005). Auswirkungen regenerativer Energiegewinnung auf die biologische Vielfalt am Beispiel der Vögel und der Fledermäuse. Bundesamt für Naturschutz, BfN-Schriften 142, Bad Godesberg.

- Hydro Tasmania. Bird and bat monitoring. Hydro Tasmania.
- Isselbächer, K. & Isselbächer, T. (2001). Vogelschutz und Windenergie in Rheinland-Pfalz, Oppenheim.
- Janss, G. (2000). Bird Behaviour In and Near a Wind Farm at Tarifa, Spain: Management Considerations. In: Proceedings of National Avian - Wind Power Planning Meeting III (ed PNAWPPM-III), pp. 110-114. Prepared for the Avian Subcommittee of the National Wind Coordinating Committee by LGL Ltd., King City, Ont., San Diego, California.
- Johnson, G.D. (2002). What is known and not known about impacts on bats? In: Proceedings of the Avian Interactions with Wind Power Structures, October 16-17, 2002 (in press), Jackson Hole, Wyoming.
- Johnson, G.D., Erickson, W.P., Strickland, D.M., Shepherd, M.F., Shepherd, D.A., & Sarappo, S.A. (2003). Mortality of Bats at a Large-scale Wind Power Development at Buffalo Ridge, Minnesota. Am. Midl. Nat., 150, 332-342.
- Johnson, G.D., Young, D.P., Erickson, W.P., Derby, C.E., Strickland, M.D., & Good, R.E. (2000). Wildlife monitoring studies Sea West Windpower Project, Carbon County, Wyoming. Western EcoSystems Technology, Inc., Cheyenne.
- Kaatz, J. (2000). Untersuchungen zur Avifauna im Bereich des Windparks Badeleben im Bördekreis - Standort- und zeitbezogene Habitatnutzung von Brut- und Rastvögeln im Prä-Post-Test-Verfahren. IHU Geologie und Analytik, Neuruppin.
- Kaatz, J., ed. (2002). Artenzusammensetzung und Dominanzverhältnisse einer Heckenbütergemeinschaft im Windfeld Nackel, pp 113-124. Technische Universität Berlin, Berlin.
- Kerlinger, P. (2000). An Assessment of the Impacts of Green Mountain Power Corporation's Searsburg, Vermont, Wind Power Facility on Breeding and Migrating Birds. In: Proceedings of National Avian - Wind Power Planning Meeting III (ed PNAWPPM-III), pp. 90-96. Prepared for the Avian Subcommittee of the National Wind Coordinating Committee by LGL Ltd., King City, Ont., San Diego, California.
- Kerns, J., Erickson, W.P., & Arnett, E.B. (2005). Bat and bird fatality at wind energy facilities in Pennsylvania and West Virginia. In: Relationship between bats and wind turbines in Pennsylvania and West Virginia: an assessment of bat fatality search protocols, patterns of fatality, and behavioral interactions with wind turbines. A final report submitted to the Bats and Wind Energy Cooperative. (ed E.B. Arnett), pp. 24-95. Bat conservation International, Austin, Texas (cited in Brinkmann & Schauer-Weissahn 2006).
- Ketzenberg, C., Exo, K.-M., Reichenbach, M., & Castor, M. (2002). Einfluss von Windenergieanlagen auf brütende Wiesenvögel. Natur und Landschaft, 77, 144-153.
- Koford, R., Jain, A., Zenner, G., & Hancock, A. (2003). Avian mortality associated with the Top of Iowa Wind Farm. Progress Report Calendar Year 2003. Iowa State University, Ames.
- Koop, B. (1997). Vogelzug und Windenergieplanung. Beispiele für Auswirkungen aus dem Kreis Plön (Schleswig-Holstein). Naturschutz und Landschaftsplanung, 29, 202-207.
- Koop, B. (1999) Windkraftanlagen und Vogelzug im Kreis Plön. Bremer Beiträge für Naturkunde und Naturschutz, 4, 25-32.
- Korn, M. & Scherner, R. (2000). Raumnutzung von Feldlerchen (*Alauda arvensis*) in einem Windpark. Natur und Landschaft, 75, 74-75.
- Kowallik, C. & Borbach-Jaene, J. (2001). Windräder als Vogelscheuchen? - Über den Einfluss der Windkraftnutzung in Gänserastgebieten an der nordwestdeutschen Küste. Vogelkundliche Berichte aus Niedersachsen, 33, 97-102.

- Kruckenberg, H. & Borbach-Jaene, J. (2001). Auswirkungen eines Windparks auf die Raumnutzung nahrungssuchender Blässhänse - Ergebnisse aus einem Monitoringprojekt mit Hinweisen auf ökoethologischen Forschungsbedarf. Vogelkundliche Berichte aus Niedersachsen, 33, 103-109.
- Kruckenberg, H. & Jaene, J. (1999). Zum Einfluss eines Windparks auf die Verteilung weidender Blässhänse im Rheiderland (Landkreis Leer, Niedersachsen). Natur und Landschaft, 74, 420-427.
- Langston, R.W.H. & Pullan, J.D. (2003). Wind farms and birds: an analysis of the effects of wind farms on birds, and guidance on environmental assessment criteria and site selection issues. Report written by BirdLife International on behalf of the Bern Convention, Sandy.
- Leddy, K.L., Higgins, K.F., & Naugle, D.E. (1999). Effects of wind turbines on upland nesting birds in Conservation Reserve Program grasslands. Wilson Bulletin, 111, 100-104.
- Lekuona, J.M. (2001). Uso del espacio por la avifauna y control de la mortalidad de aves y murciélagos en los parques eólicos de Navarra durante un ciclo anual. Dirección General de Medio Ambiente, Gobierno de Navarra, Pamplona.
- Lucas, M.d., Janss, G.F.E., & Ferrer, M. (2005). A bird and small mammal BACI and IG design studies in a wind farm in Malpica (Spain). Biodiversity and Conservation, 14, 3289-3303.
- Manville, A.M. (2001). Communication Towers, Wind Generators, and Research: Avian Conservation Concerns. In: Proceedings of National Avian - Wind Power Planning Meeting IV (ed PNAWPPM-IV), pp. 152-159. Prepared for the Avian Subcommittee of the National Wind Coordinating Committee by RESOLVE, Inc., Washington, D.C., Susan Savitt Schwartz, Carmel, California.
- Meek, E.R., Ribbands, J.B., Christer, W.G., Davey, P.R., & Higginson, I. (1993). The effects of aero-generators on moorland bird populations in the Orkney Islands, Scotland. Bird Study, 40, 140-143.
- Menzel, C. (2002). Rebhuhn und Rabenkrähe im Bereich von Windkraftanlagen im niedersächsischen Binnenland, pp 97-112. Technische Universität Berlin, Berlin.
- Menzel, C. & Pohlmeier, K. (1999). Indirekter Raumnutzungsnachweis verschiedener Niederwildarten mit Hilfe von Lösungsstangen („dropping marker“) in Gebieten mit Windkraftanlagen. Z. Jagdwiss., 45, 223-229.
- Musters, C.J.M., Noordervliet, M.A.W., & Keurs, W.J.T. (1996). Bird casualties caused by a wind energy project in an estuary. Bird Study, 43, 124-126.
- Orloff, S. & Flannery, A. (1996). A continued examination of avian mortality in the Altamont Pass Wind Resource Area. California Energy Commission, Sacramento; Bio-Systems Analysis, Inc., Santa Cruz, California.
- Osborn, R.G., Higgins, K.F., Dieter, C.D., & Usgaard, R.E. (1996). Bat collisions with wind turbines in Southwest Minnesota. Bat Research News, 37, 105-108.
- Pedersen, M.B. & Poulsen, E. (1991). Impact of a 90m/2 MW wind turbine on birds. Avian responses to the implementation of the Tjaereborg Wind Turbine at the Danish Wadden Sea. Dansk Vildtundersøgelse Kalø, 47.
- Percival, S.M. (2000). Birds and wind turbines in Britain. British Wildlife, 12, 8-15.
- Petersen, I.K., Clausager, I., & Christensen, T.K. (2004). Bird numbers and distribution in the Horns Rev offshore wind farm area. NERI - report.
- Phillips, J.F. (1994). The effects of a windfarm on the upland breeding bird communities of Bryn Titli, Mid Wales: 1993-1994. Royal Society for the Protection of Birds, The Welsh Office, Bryn Aderyn, The Bank, Newtown, Powys.

- Reichenbach, M. (2002). Windenergie und Wiesenvögel - wie empfindlich sind die Offenlandbrüter?, pp 52-76. Technische Universität Berlin, Berlin.
- Reichenbach, M. (2003a). Auswirkungen von Windenergieanlagen auf Vögel - Ausmaß und planerische Bewältigung, Technische Universität Berlin, Berlin.
- Reichenbach, M. (2003b). Windenergie und Vögel - Ausmaß und planerische Bewältigung. Dissertation, Technische Universität, Berlin.
- Reichenbach, M. & Schadek, U. (2003). Langzeituntersuchungen zum Konfliktthema „Windkraft und Vögel“. 2. Zwischenbericht. Unveröffentlichtes Gutachten im Auftrag des Bundesverbandes Windenergie.
- Reichenbach, M. & Sinning, F. (2003). Empfindlichkeiten ausgewählter Vogelarten gegenüber Windenergieanlagen - Ausmaß und planerische Bewältigung. In: Kommen die Vögel und Fledermäuse unter die (Wind)räder?, Dresden, 17.-18.11.2003.
- Reichenbach, M. & Steinborn, H. (2006). Langzeituntersuchungen zum Konfliktthema „Windkraft und Vögel“. ARSU GmbH, Oldenburg.
- Sachslehner, L. & Kollar, H.P. (1997). Vogelschutz und Windkraftanlagen in Wien. Stadt Wien, Wien.
- Scherner, E.R. (1999). Windkraftanlagen und „wertgebende Vogelbestände“ bei Bremerhaven: Realität und Realsatire? Beiträge zur Naturkunde Niedersachsens, 52, 121-156.
- Schmidt, E., Piaggio, A.J., Bock, C.E., & Armstrong, D.M. (2003). National Wind Technology Center Site Environmental Assessment: Bird and Bat Use and Fatalities - Final Report; Period of Performance: April 23, 2001 - December 31, 2002. NREL/SR-500-32981.
- Schreiber, M. (1992). Rastvögel und deren Habitatwahl im Bereich „Westermarsch“ (Landkreis Aurich) im Jahr 1992. Unveröff. Gutachten im Auftrag der Ingenieursgemeinschaft agwa.
- Schreiber, M. (1993a). Windkraftanlagen und Watvogel-Rastplätze - Störungen und Rastplatzwahl von Brachvogel und Goldregenpfeifer. Naturschutz und Landschaftsplanung, 25, 133-139.
- Schreiber, M. (1993b). Zum Einfluß von Störungen auf die Rastplatzwahl von Watvögeln. Informationsd. Natursch. Nieders., 13, 161-169.
- Schreiber, M. (1999). Windkraftanlagen als Störungsquelle für Gastvögel am Beispiel von Blessgans (*Anser albifrons*) und Lachmöwe (*Larus ridibundus*). Bremer Beiträge für Naturkunde und Naturschutz, 4, 39-48.
- Schreiber, M. (2002). Einfluss von Windenergieanlagen auf Rastvögel und Konsequenzen für EU-Vogelschutzgebiete, Technische Universität Berlin, Berlin.
- SEO (1995). Effects of wind turbine power plants on the Avifauna in the Campo de Gibraltar region, Rep. No. Report to the Environmental Agency. Sociedad Espanola de Ornitologia SEO.
- SGS Environment (1994). Haverigg windfarm ornithological monitoring programme. Report to Windcluster LTD.
- Sinning, F. (1999). Ergebnisse von Brut- und Rastvogeluntersuchungen im Bereich des Jade-Windparks und DEWI-Testfelds in Wilhelmshaven. Bremer Beiträge für Naturkunde und Naturschutz, 4, 61-70.
- Sinning, F. (2004a). Bestandsentwicklung von Kiebitz (*Vanellus vanellus*), Rebhuhn, (*Perdix perdix*) und Wachtel (*Coturnix coturnix*) im Windpark Lahn (Niedersachsen, Lkr. Emsland) - Ergebnisse einer 6-jährigen Untersuchung. Bremer Beiträge für Naturkunde und Naturschutz, 7, 97-103.

- Sinning, F. (2004b). Kurzbeitrag zum Vorkommen der Grauammer (*Miliaria calandra*) und weiterer ausgewählter Arten an Gehölzreihen im Windpark Mallnow (Brandenburg, Landkreis Märkisch Oderland). Bremer Beiträge für Naturkunde und Naturschutz, 7, 193-196.
- Sinning, F. (2004c). Kurzbeitrag zum Vorkommen des Schwarzkehlchens (*Saxicola torquata*) und weiterer ausgewählter Arten in zwei norddeutschen Windparks (Niedersachsen, Landkreise Ammerland, Leer und Stade). Bremer Beiträge für Naturkunde und Naturschutz, 7, 199-203.
- Sinning, F. & Bruyn, U.d. (2004). Raumnutzung eines Windparks durch die Vögel während der Zugzeit - Ergebnisse einer Zugvogel- Untersuchung im Windpark Wehrder (Niedersachsen, Landkreis Wesermarsch). Bremer Beiträge für Naturkunde und Naturschutz, 7, 157-179.
- Sinning, F. & Gerjets, D. (1999). Untersuchungen zu Annäherung rastender Vögel in Windparks in Nordwestdeutschland. Bremer Beiträge für Naturkunde und Naturschutz, 4, 53-59.
- Sinning, F., Sprötge, M., & Bruyn, U.d. (2004). Veränderung der Brut-und Rastvogelfauna nach Errichtung des Windparks Abens-Nord (Niedersachsen,Landkreis Wittmund). Bremer Beiträge für Naturkunde und Naturschutz, 7, 77-91.
- Smallwood, K.S. & Thelander, C.G. (2004). Developing methods to reduce bird mortality in the Altamont Pass Wind Ressource Area. Final report by BioResource Consultants to the California Energy Commission.
- Sommerhage, M. (1997). Verhaltensweisen ausgesuchter Vogelarten gegenüber Windkraftanlagen auf der Vaßbecker Hochfläche (Landkreis Waldeck-Frankenberg). Vogelkundliche Hefte Edertal, 23, 104-109.
- Steiof, K., Becker, J., & Rathgeber, J. (2002). Ornithologische Stellungnahme zur Erweiterung der Windenergieanlage bei Mildenberg (Kreis Oberhavel, Land Brandenburg). Gutachten im Auftrag der Windenergie Wenger-Rosenau GmbH, Berlin.
- Still, D., Little, B., & Lawrence, E.S. (1996). The effect of wind turbines on the bird population at Blyth Harbour, Northumberland. ETSU W/13/00394/REP.
- Strickland, M.D., Johnson, G., Erickson, W.P., & Kronner, K. (2001). Avian Studies at Wind Plants Located at Buffalo Ridge, Minnesota and Vansycle Ridge, Oregon. In: Proceedings of National Avian - Wind Power Planning Meeting IV (ed PNAWPPM-IV), pp. 38-52. Prepared for the Avian Subcommittee of the National Wind Coordinating Committee by RESOLVE, Inc., Washington, D.C., Susan Savitt Schwartz, Carmel, California.
- Stübing, S. & Bohle, H.W. (2001). Untersuchungen zum Einfluss von Windenergieanlagen auf Brutvögel im Vogelsberg (Mittelhessen). Vogelkundliche Berichte aus Niedersachsen, 33, 111-118.
- Thelander, C.G. & Rugge, L. (2000). Avian risk behavior and fatalities at the Altamont Wind Resource Area, March 1998 to February 1999. NREL/SR-500-27545.
- Thelander, C.G., Smallwood, K.S., & Rugge, L. (2003). Bird risk behaviors and fatalities at the Altamont Pass Wind Resource Area. Period of performance: March 1998 - December 2000. NREL/SR-500-33829.
- Trapp, H., Fabian, D., Förster, F., & Zinke, O. (2002). Fledermausverluste in einem Windpark. Naturschutzarbeit in Sachsen, 44, 53-56.
- Traxler, A., Wegleiter, S., & Jaklitsch, H. (2004). Vogelschlag, Meideverhalten & Habitatnutzung an bestehenden Windkraftanlagen Prellenkirchen - Obersdorf - Steinberg/ Prinzendorf. Biome, Gerasdorf.

- Ugoretz, S. (2001). Avian Mortalities at Tall Structures. In: Proceedings of National Avian - Wind Power Planning Meeting IV (ed PNAWPPM-IV), pp. 165-166. Prepared for the Avian Subcommittee of the National Wind Coordinating Committee by RESOLVE, Inc., Washington, D.C., Susan Savitt Schwartz, Carmel, California.
- van der Winden, J., Spaans, A.L., & Dirksen, S. (1999). Nocturnal collision risks of local wintering birds with wind turbines in wetlands. Bremer Beiträge für Naturkunde und Naturschutz, 4, 33-38.
- Vierhaus, H. (2000). Neues von unseren Fledermäusen. ABU Info, 24, 58-60.
- Walter, G. & Brux, H. (1999). Erste Ergebnisse eines dreijährigen Brut- und Gastvogelmonitorings (1994-1997) im Einzugsbereich von zwei Windparks im Landkreis Cuxhaven. Bremer Beiträge für Naturkunde und Naturschutz, 4, 81-106.
- Winkelman, J.E. (1989). Vogels in het windpark nabij Urk (NOP): aanvaringsslachtoffers en verstoring van pleisterende eenden, ganzen en zwanen. RIN-rapport 89/15, Arnhem.
- Winkelman, J.E. (1992a). De invloed van de Seproef Windcentrale te Oosterbierum (Fr.) op vogels. 1: aanvaringsslachtoffers. DLO Instituut voor Bos- en Natuuronderzoek (Hrsg.), Arnhem.
- Winkelman, J.E. (1992b). De invloed van de Sep-proefwindcentrale te Oosterbierum (Fr.) op vogels, 4: verstoring. RIN-rapport92/5, Arnhem.
- Young, D.P., Erickson, W.P., Good, R.E., Strickland, M.D., & Johnson, G.D. (2003a). Avian and bat mortality associated with the initial phase of the Foote Creek Rim Wind-power project, Carbon County, Wyoming. Final report. Western EcoSystems Technology, Inc., Wyoming.
- Young, D.P., Erickson, W.P., Strickland, M.D., Good, R.E., & Sernka, K.J. (2003b). Comparison of avian responses to UV-light-reflective paint on wind turbines. Western EcoSystems Technology, NREL/SR-500-32840, Cheyenne.

10 Appendix

Bird casualties at wind turbines in Germany and casualties of bats at wind turbines worldwide. Total of findings since 1989 (intensive search since 2002). Data from the central card index of the state bird protection office in the Landesumweltamt (state environment department) Brandenburg. Compiled by : Tobias Dürr, as of 16.10.2006.

Bird casualties at wind turbines in Germany													
Data from the central card index of the state bird protection office													
In the Landesumweltamt (state environment department) Brandenburg													
Compiled by: Tobias Dürr; as of: 16. October 2006													
State/Land													Tot.
Species		BB	ST	SN	TH	MV	SH	NI	HB	NW	HE	SL	BW
Gavia stellata	Red-throated Diver								1				1
Phalacrocorax carbo	Cormorant								2				2
Ciconia ciconia	White Stork	4				3	1					1	9
Ciconia nigra	Black Stork									1			1
Cygnus cygnus	Whooper Swan						1						1
Cygnus olor	Mute Swan	2	1			1	1	5					10
Anser anser	Greylag Goose							1					1
Anser fabalis	Bean Goose			1									1
Anser albifrons	White-fronted Goose		1										1
Anser fabalis / albifrons	Bean-/White-fronted Goose	1	1										2
Branta leucopsis	Barnacle Goose					6							6
Tadorna tadorna	Shelduck								1				1
Anas crecca	Teal								1				1
Anas platyrhynchos	Mallard	3		1			6	1	2				13
Anas clypeata	Shoveler						1						1
Aythya fuligula	Tufted Duck							1					1
Haliaeetus albicilla	White-tailed Eagle	4	2			9	8	1					24
Milvus milvus	Red Kite	33	22	8	3	1		2		1	6		76
Milvus migrans	Black Kite	6		1									7
Accipiter gentilis	Goshawk	2		1									3
Accipiter nisus	Sparrowhawk	1		1									2
Buteo buteo	Common Buzzard	37	14	3	2		1	4		1	1	1	64
Buteo lagopus	Rough-legged Buzzard		1										1
Circus aeruginosus	Marsh Harrier	1											1
Circus pygargus	Montagu's Harrier								1				1
Falco subbuteo	Hobby	2											2
Falco columbarius	Merlin	1											1
Falco tinnunculus	Kestrel	7	7	1					1				16
<i>Falconiformes spp.</i>	<i>Raptor spp..</i>	1											1
Perdix perdix	Grey Partridge	1											1
Phasianus colchicus	Pheasant	1	1					1	1				4
Fulica atra	Coot				3	1							4
Haematopus ostralegus	Oystercatcher					2	1						3
Pluvialis apricaria	Golden Plover		2			8							10
Vanellus vanellus	Northern Lapwing					3							3
Gallinago gallinago	Common Snipe					1							1
Larus ridibundus	Black-headed Gull	4					12	2	2				20
Larus argentatus	Herring Gull						12	2	1				15
Larus fuscus	Lesser Black-backed Gull							1					1
Larus canus	Common Gull	3				6	2	2					13
Chlidonias niger	Black Tern						1						1
Uria aalge	Guillemot							1					1
<i>Columba livia f. domestica</i>	Feral Pigeon	14		1		2							17
Columba oenas	Stock Dove	3											3
Columba palumbus	Wood Pigeon	14	2										16
Tyto alba	Barn Owl	1											1
Asio otus	Long-eared Owl	1											1
Asio flammea	Short-eared Owl	2											2
Bubo bubo	Eagle Owl			2					3		1	6	
Cuculus canorus	Cuckoo	1											1
Apus apus	Swift	11	2	1					1		2	17	
Apus melba	Alpine Swift										1		1
<i>Apus spec.</i>	<i>Swift sp.</i>										1		1
Picus viridis	Green Woodpecker	1											1
Dendrocopos major	Great Spotted Woodpecker	1											1
<i>Nonpasseriformes spec.</i>		1											1

Species		State/Land												Tot.
		BB	ST	SN	TH	MV	SH	NI	HB	NW	HE	SL	BW	
<i>Alauda arvensis</i>	Skylark	19			3									22
<i>Lullula arborea</i>	Woodlark	1												1
<i>Eremophila alpestris</i>	Shorelark						1							1
<i>Anthus trivialis</i>	Tree Pipit	1												1
<i>Hirundo rustica</i>	Barn Swallow	3					1							4
<i>Delichon urbica</i>	House Martin	3										3		6
<i>Motacilla alba</i>	White Wagtail	1												1
<i>Motacilla flava</i>	Yellow Wagtail	1												1
<i>Troglodytes troglodytes</i>	Wren	1												1
<i>Acrocephalus palustris</i>	Sedge Warbler						1							1
<i>Hippolais polyglotta</i>	Melodious Warbler											1		1
<i>Sylvia curruca</i>	Lesser Whitethroat	1												1
<i>Sylvia atricapilla</i>	Common Whitethroat	2												2
<i>Regulus regulus</i>	Goldcrest	2		1						1				4
<i>Regulus ignicollis</i>	Firecrest	1												1
<i>Regulus spec.</i>	Goldcrest/Firecrest											1		1
<i>Ficedula hypoleuca</i>	Pied Flycatcher	3												3
<i>Saxicola rubetra</i>	Whinchat	1												1
<i>Erythacus rubecula</i>	Robin	2												2
<i>Turdus pilaris</i>	Fieldfare	1	1											2
<i>Turdus philomelos</i>	Song Thrush	1												1
<i>Turdus iliacus</i>	Redwing						1							1
<i>Turdus merula</i>	Blackbird	2	1											3
<i>Parus major</i>	Great Tit	1												1
<i>Emberiza calandra</i>	Corn Bunting	13												13
<i>Emberiza citrinella</i>	Yellowhammer	6		1										7
<i>Carduelis chloris</i>	Greenfinch	2												2
<i>Carduelis flavirostris</i>	Twite						1							1
<i>Fringilla coelebs</i>	Chaffinch	2	1	1	1									4
<i>Passer montanus</i>	Tree Sparrow	2												2
<i>Passer domesticus</i>	House Sparrow	1												1
<i>Sturnus vulgaris</i>	Starling	11	1	1	1		3							17
<i>Lanius collurio</i>	Red-backed Shrike	1												1
<i>Garrulus glandarius</i>	Jay	2												2
<i>Pica pica</i>	Magpie		1											1
<i>Corvus corax</i>	Raven	10												10
<i>Corvus frugilegus</i>	Rook		1											1
<i>Corvus corone</i>	Carriion/Hooded crow	6									1			7
<i>Corvus spec.</i>	Crow spp.	1						1						2
		257	61	20	14	17	80	26	14	9	9	1	11	517

BB = Brandenburg, ST = Sachsen-Anhalt, SN = Sachsen, TH = Thüringen, MVP = Mecklenburg-Vorpommern,
SH = Schleswig-Holstein, NI = Niedersachsen, HB = Hansestadt Bremen, NW = Nordrhein-Westfalen, HE = Hessen,
SL = Saarland, BW = Baden-Württemberg

Bat casualties at wind turbines

Compiled by: Tobias Dürr, Landesumweltamt Brandenburg -Ref. Ö2 / Staatliche Vogelschutzwarte, Buckower Dorfstraße 34, D-14715 Nennhau
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Species		BB	ST	SN	MV	SH	N	NW	RP	HE	BW	SL	BY	D ges.	ESP	POR	A	FRA	SWE	USA	AUS	ges.	
<i>Nyctalus noctula</i>	Noctule	110	2	29	4	6	5	3				1	160	1		3	3	1			168	1	
<i>Nyctalus lasiopterus</i>	Greater Noctule												0	1									
<i>N. leisleri</i>	Leisler's Bat	9	2	3	1		3					16		34	1							35	
<i>Eptesicus serotinus</i>	Serotine	6	3		1	2			1		1		13	1								14	
<i>E. nilssonii</i>	Northern Bat	1										1										9	
<i>Vesperugo discolor</i>	Parti-coloured Bat	7	7	5			1	2			22								1			23	
<i>Myotis myotis</i>	Mouse-eared Bat										0	1										1	
<i>M. dasycneme</i>	Pond Bat				1						1											1	
<i>M. daubentonii</i>	Daubenton's Bat	1		1	1						3											3	
<i>M. brandtii</i>	Brandt's Bat	1									1											1	
<i>Pipistrellus pipistrellus</i>	Common Pipistrelle	22	2	13	4	6	1	3		101		152	1					2	1			156	
<i>P. nathusii</i>	Nathusius's Pipistrelle	65	4	33	11	1	9	1	1		126		1	30	5							162	
<i>P. pygmaeus</i>	Soprano Pipistrelle	6		1							7			1								8	
<i>P. kuhlii</i>	Kuhl's Pipistrelle										0	1										1	
<i>Pipistrellus sp.</i>	Pipistrellus sp.	3		1			4				8			4								12	
<i>Hypsugo savii</i>	Savi's Bat										0	3										3	
<i>Plecotus austriacus</i>	Grey Long-eared Bat	5	1								6		1									7	
<i>Plecotus auritus</i>	Brown Long-eared Bat			1	1						2											2	
<i>Tadarida teniotis</i>	Bulldog Bat										0	1										1	
<i>Miniopterus schreibersii</i>	Schreiber's Long-fingered Bat										0	1										1	
<i>Chiroptera sp.</i>	Bat spez.	2	2	2	1		2	1		10	14			1			1	30	60		115		
<i>Lasiurus cinereus</i>	Hoary Bat																				637	637	
<i>L. borealis</i>	Red Bat																				341	341	
<i>Lasionycteris noctivagans</i>	Silver-haired Bat																				96	96	
<i>Eptesicus fuscus</i>	Big brown Bat																				96	96	
spec. ?	Southern Brown Bat																				22	22	
<i>Myotis luciferus</i>	Little Brown Bat																				88	88	
<i>M. septentrionalis</i>	Northern Long-eared																				6	6	
<i>M. evotis</i>	Long Eared																				1	1	
<i>M. spec.</i>	Myotis sp.																	2	2		2	2	
<i>Pipistrellus subflavus</i>	Eastern Pipistrelle																				108	108	
<i>Tadarida brasiliensis</i>	Mexican Free-tailed Bat																	1			1	1	
<i>Tadarida australis</i>	Australian Free-tailed Bat																				29	29	
<i>Chalinolobus mario</i>	Chocolate-Wattled Bat	236	13	92	28	9	24	2	12	2	2	125	0	1	546	25	1	5	40	47	1416	29	2110

BB = Brandenburg, ST = Sachsen-Anhalt, SN = Sachsen, TH = Thüringen, MV = Mecklenburg-Vorpommern, SH = Schleswig-Holstein, NI = Niedersachsen, NW = Nordrhein-Westfalen, RP = Rheinland-Pfalz, HE = Hessen, BW = Baden-Württemberg, SL = Saarland, BY = Bayern, POR = Spain, ESP = Spain, A = Austria, FRA = France, SWE = Sweden, USA = United States of America, AUS = Australia